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2 **Retrieving Particulate Matter Concentrations over the Contiguous United States**
3 **Using CALIOP Observations**
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7 Travis D. Toth¹, Jianglong Zhang², Mark A. Vaughan¹, Jeffrey S. Reid³, and James R. Campbell³
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11 ¹NASA Langley Research Center, Hampton, VA

12 ²University of North Dakota, Grand Forks, ND

13 ³Naval Research Laboratory, Monterey, CA
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45 Corresponding Author: travis.d.toth@nasa.gov

Abstract

Using twelve years (2007-2018) of NASA Cloud-Aerosol Lidar with Orthogonal Polarization (CALIOP) near-surface 532 nm aerosol extinction retrievals, multi-year mean and trends of particulate matter (PM) concentrations are derived over the contiguous United States (CONUS). Different from past studies that use column integrated aerosol optical thickness, here only near-surface CALIOP aerosol extinction is used for deriving near-surface PM with aerodynamic diameters less than $2.5\text{ }\mu\text{m}$ ($\text{PM}_{2.5}$) concentrations using an innovative, bulk-mass-modeling-based method. Compared against ground based $\text{PM}_{2.5}$ measurements from the U.S. Environmental Protection Agency (EPA), an encouraging relationship between CALIOP-derived $\text{PM}_{2.5}$ and EPA-observed $\text{PM}_{2.5}$ (Deming slope = 0.89; RMSE = $3.42\text{ }\mu\text{g}/\text{m}^3$; mean bias = $-1.00\text{ }\mu\text{g}/\text{m}^3$) is found using combined daytime/nighttime CALIOP data. Also, comparable trends in $\text{PM}_{2.5}$ concentrations from the EPA and daytime and nighttime CALIOP data are found for most of the eastern CONUS and imply that air quality is generally improving over this region for the study period. Over the western CONUS, a seasonal analysis reveals that $\text{PM}_{2.5}$ trends are positive during the more active wildfire season (June through November) but negative for other months. This study suggests that lidar data show promise in their use for obtaining $\text{PM}_{2.5}$ estimates and provides motivation to further explore aerosol extinction-based PM concentration retrievals in anticipation of future space-based lidar missions.

1 Introduction

Aerosol particles, especially particulate matter (PM) with aerodynamic diameters smaller than 2.5 μm ($\text{PM}_{2.5}$), contribute to air pollution and negatively impact human health (e.g., Schwartz et al., 1996; Pope et al., 2002; Xing et al., 2016). As such, the United States (U.S.) Environmental Protection Agency (EPA) continually monitors $\text{PM}_{2.5}$ concentrations across the country through a ground-based network of *in situ* instruments to support air quality forecasting and decision making for environment-related policies (e.g., Padgett and Richmond, 1983; Federal Register, 1997). However, while the U.S., as well as members of the European Union, have extensive monitoring networks with freely available data, measurements are more limited over many areas of the world. In response, researchers have investigated the use of aerosol optical thickness (AOT) derived from space-based passive remote sensing instruments for $\text{PM}_{2.5}$ applications (e.g., Chu et al., 2003; Wang and Christopher, 2003; Van Donkelaar et al., 2006; Lee et al., 2012; Xie et al., 2015). These past efforts have been primarily based upon correlative relationships between EPA $\text{PM}_{2.5}$ and satellite-based AOT estimates (e.g., Hoff and Christopher, 2009 and references therein), with some studies leveraging chemical transport models in an attempt to improve the $\text{PM}_{2.5}$ /AOT relationship (e.g., Liu et al., 2004; Van Donkelaar et al., 2016).

The clear advantage of the passive remote sensor AOT approach to estimating $\text{PM}_{2.5}$ comes from the large spatial and temporal coverages that satellites provide. However, $\text{PM}_{2.5}$ concentrations are surface-based measurements in units of micrograms per cubic meter, whereas AOT is a unitless column-integrated property that measures the amount of attenuation of solar energy due to aerosols through the full atmospheric column. Thus, for elevated aerosol plumes above the planetary boundary layer (PBL) and near-ground surface layer effects, $\text{PM}_{2.5}$ and AOT can be uncorrelated (e.g., Toth et al., 2014; Reid et al., 2017). Also, while AOT provides a single,

column-integrated estimate of aerosol loading, vertical variations in particle size and hygroscopic growth within a column can be large and are often non-uniformly distributed. Therefore, AOT is not necessarily a reliable proxy for surface $\text{PM}_{2.5}$ concentrations in many cases (e.g., Hand and Malm, 2007; Toth et al., 2014; Kaku et al., 2018).

An alternate approach to using radiometric proxy estimates derived from passive remote sensors is to directly derive $\text{PM}_{2.5}$ concentrations using aerosol extinction profiled near the Earth's surface using range-resolved measurements from space-based lidars. The advantage here is that discrete near-surface extinction coefficients can more accurately capture aerosol optical properties at the surface than an integrated parameter such as AOT. In a recent proof-of-concept study, Toth et al. (2019) describe a bulk-mass-modeling method to directly retrieve $\text{PM}_{2.5}$ using near-surface Cloud-Aerosol Lidar with Orthogonal Polarization (CALIOP) observations over the contiguous United States (CONUS) for a two-year (2008-2009) period. In this algorithm, $\text{PM}_{2.5}$ concentration or aerosol mass concentration is derived by dividing the CALIOP retrieved near-surface extinction coefficient by the product of the aerosol mass extinction efficiency (e.g., Liou, 2002; Chow et al., 2006; Hand and Malm, 2007), the hygroscopic growth of particles, and inverse of the $\text{PM}_{2.5}$ to PM_{10} (particulate matter with aerodynamic diameters smaller than $10\text{ }\mu\text{m}$) conversion ratio. The analyses resulted in R^2 values between EPA $\text{PM}_{2.5}$ and CALIOP-derived $\text{PM}_{2.5}$ ranging from ~ 0.2 for daytime CALIOP observations to ~ 0.5 for nighttime observations.

We now examine the feasibility of applying the algorithm described in Toth et al. (2019) to nearly the entire data record of CALIOP (2007-2018) and study the long-term means and trends of the derived $\text{PM}_{2.5}$ concentration over the CONUS. We restrict the analysis to the CONUS to maximize our opportunities for comparisons with the very large number of well-validated ground-based $\text{PM}_{2.5}$ measurements acquired by the EPA. We investigate the regional variability of EPA

and CALIOP-based $PM_{2.5}$ concentrations over the CONUS and describe sensitivity studies of assumed mixed layer height (defined as those near-surface CALIOP range bins used for deriving $PM_{2.5}$) and. Because traditional AOT/ $PM_{2.5}$ studies frequently use full column AOT as a proxy for near-surface aerosol extinction (e.g., Hoff and Christopher, 2009), in this work we report the mean state and trends in the contribution of near-surface aerosols to those of the total atmospheric column. We focus on the following research questions:

1. How consistent is the Toth et al. (2019) algorithm when extended over twelve years of CALIOP data?
2. Do any trends in CALIOP-derived $PM_{2.5}$ over the CONUS appear in the data, and how do they compare with those from EPA?
3. Are there regional and/or seasonal differences in the mean state and trend of EPA versus CALIOP-derived $PM_{2.5}$ concentrations throughout the CONUS during the study period?
4. How sensitive are the EPA/CALIOP $PM_{2.5}$ correlations to the assumed mixed layer height?
5. How does the fraction of near-surface AOT to column AOT vary over the study time period?

The primary goal of this research is to address the 2017 NASA Decadal Survey that prioritizes the need for an increased understanding of air pollution distribution and its trends (National Academies, 2018) and lays the foundation for the future characterization of air pollution from the next lidars in space (e.g., NASA Atmosphere Observing System; Stephens et al., 2021). The $PM_{2.5}$ analyses provided in this paper have the potential for use in air quality research, applications, and model validation, helping to quantify $PM_{2.5}$ concentrations over areas with few

to no surface stations, and setting the stage for the development of a more robust CALIOP-based PM_{2.5} product with possible implementation on a global scale.

2 Data, Methods, and Algorithm

2.1 EPA data

The EPA operates an extensive network of ground stations across the country that routinely monitor atmospheric components responsible for degrading regional air quality, including criteria gases (ozone, SO₂, CO, and NO₂) and particulate matter (PM_{2.5} and PM₁₀), providing data at daily and hourly resolutions (U.S. EPA, 2020). PM_{2.5} measurements are acquired through a variety of instruments that adhere to Federal Reference Method (FRM; gravimetric analysis) and Federal Equivalent Method (FEM; taper element oscillating microbalance [TEOM] and beta gauge analyses) regulations (Federal Register, 1997; Noble et al., 2001; Greenstone, 2002). In this study, twelve years (2007-2018) of Daily PM_{2.5} Local Conditions data (parameter code 88101) from EPA stations are used for analysis, obtained from the EPA Air Quality System (AQS). This dataset is comprised mostly of 24-hour filter-based (i.e., gravimetric) measurements, but also includes an average of hourly TEOM or beta gauge measurements over a 24-hour period for some stations. Uncertainties in EPA PM_{2.5} data are instrument/method dependent, as summarized in Toth et al. (2019), and explanations of PM_{2.5} uncertainties in a greater level of detail can be found in other studies (e.g., Spagnolo, 1989; Chung et al., 2001; Patashnick et al., 2001; Eatough et al, 2003; Kiss et al., 2017).

2.2 CALIPSO data

Operational since 2006, CALIOP is a dual-wavelength (532 nm and 1064 nm) polarization-sensitive (at 532 nm) elastic backscatter lidar flying aboard the Cloud-Aerosol Lidar and Infrared Pathfinder Satellite Observations (CALIPSO) satellite, providing a vertically-resolved perspective of aerosols and clouds in the atmosphere to the limit of signal attenuation (Winker et al., 2010). The range-resolved aerosol measurements made by CALIOP have been applied for air quality studies, including those involving PM_{2.5} (e.g., Toth et al., 2014; 2019; Bin et al., 2021; Fang et al., 2021) and detection/movement of pollution episodes (e.g., Tao et al., 2012; Kar et al., 2015; Chen et al., 2018; Yin et al., 2021). Initially a member of the “A-Train” satellite constellation (Stephens et al., 2018), CALIPSO joined CloudSat in its orbit in late 2018 to form the “C-Train” satellite constellation (flying ~16.5 km below “A-Train”; e.g., Yeom et al., 2020).

In this study, twelve years (2007-2018) of 532 nm extinction coefficient from the V4.10 CALIOP Level 2 5 km aerosol profile (L2_05kmAPro) product are analyzed over the CONUS during both daytime and nighttime conditions, respectively. The ending year of 2018 was chosen due to CALIPSO’s new orbit in the “C-Train” after that time and the desire to have consistent sampling for the trend analysis. The method used here for processing CALIPSO aerosol data is consistent with several previous papers, all of which provide detailed explanations (Toth et al., 2014; 2016; 2019). In a general sense, the CALIPSO data are first subject to rigorous quality assurance (QA) and cloud screening procedures, similar to those described in Kittaka et al. (2011), Campbell et al. (2012), and Winker et al. (2013), through the use of parameters from the L2_05kmAPro and CALIOP Level 2 5 km aerosol layer (L2_05kmALay) products. The specific QA parameters and thresholds used in this study are outlined in Toth et al. (2019). The aerosol profiles are then linearly re-gridded from 60 m vertical resolution (above mean sea level; AMSL) to 100 m bins, referenced to the local surface (above ground level; AGL). The means of aerosol

extinction are then computed for the 100-1000 m AGL altitude range, and this is defined as near-surface aerosol extinction for this study. The lowest 100 m bin is not considered due to potential surface contamination in aerosol extinction (e.g., Kim et al., 2017; Toth et al., 2019), but sensitivity studies concerning this topic are explored in Sect. 3.3. For later validation purposes, CALIOP data are spatially (within 100 km) and temporally (same day) collocated with data from the EPA (consistent with Toth et al., 2019).

2.3 Methods and Algorithm

As described in Toth et al. (2019), the algorithm to derive PM mass concentration ($\mu\text{g}/\text{m}^3$) is based on the normalization of the 532 nm extinction coefficient (km^{-1}) by the mass extinction (scattering + absorption) efficiency (m^2/g ; Liou, 2002; Chow et al., 2006), written as

$$C_m = \frac{\sigma \times 1000}{(\alpha_{scat} \times f_{rh} + \alpha_{abs})}, \quad (1)$$

where σ is the CALIOP-derived mean near-surface (100-1000 m AGL) aerosol extinction coefficient, α_{scat} and α_{abs} are the dry mass scattering and absorption efficiencies, respectively, f_{rh} is the hygroscopic growth factor, and C_m is the PM mass concentration. Pollution particle composition is assumed to be essentially uniform over all of the CONUS, and hence the corresponding α_{scat} and α_{abs} values used are $3.40 \text{ m}^2/\text{g}$ and $0.37 \text{ m}^2/\text{g}$, respectively. These values were obtained from the Optical Properties of Aerosols and Clouds (OPAC) model (Hess et al., 1998) and are consistent with previous studies (e.g., Hand and Malm, 2007; Kaku et al., 2018). We recognize that the assumption of one aerosol type over the entire CONUS is a significant simplification. A future study is planned to fully investigate the effect of different aerosol types

(and subsequent mass scattering/absorption efficiencies) on extinction-based retrievals of PM_{2.5} concentrations, similar in principle to the analysis of Omar et al. (2005) that develops an optical model for six modular global aerosol types that served as a predicate to the original species distinguished by CALIOP algorithms (Omar et al., 2009).

Also, note that in order to focus on fine mode aerosols, we exclude extinction range bins classified as pure dust (which exhibit a_{scat} values $\sim 0.5\text{-}0.7$ m²/g; e.g., Malm and Hand, 2007) by the CALIPSO aerosol typing algorithms. We acknowledge, however, that dust is not coarse mode alone (e.g., Omar et al., 2005), and some fine mode dust will be present in the EPA-observed PM_{2.5} concentrations. This is particularly true over parts of the western CONUS, where dust aerosols are generally more prevalent compared to other regions in the CONUS (e.g., Omar et al., 2009). Thus, our choice to remove CALIOP-classified dust profiles may impact the performance of our PM_{2.5} derivations when comparing to ground station data, and also lowers the number of CALIPSO points available for analysis (e.g., especially over the western CONUS, as discussed further in Sect. 3). The implications of removing dust will be fully explored in a future study in CALIOP-derived PM_{2.5} concentrations as a function of aerosol type.

Furthermore, the hygroscopic growth factor is a necessary component of Eq. (1) because *in situ* PM_{2.5} are dry mass measurements, but extinction retrievals consider aerosol humidification effects. This factor is computed from Modern-Era Retrospective analysis for Research and Applications Version 2 (MERRA-2) relative humidity (RH) profiles included in the CALIPSO datasets using the approach by Hänel (1976), as

$$f_{rh} = \left(\frac{1 - RH}{1 - RH_{ref}} \right)^{\Gamma}, \quad (2)$$

where RH_{ref} is the reference RH (30% for this analysis; Lynch et al., 2016) and Γ is a unitless light scattering hygroscopicity fit parameter (0.63 for this study, assuming sulfate aerosol; e.g., Toth et al., 2019). A $PM_{2.5}$ to PM_{10} ratio (ϕ) is also considered in the $PM_{2.5}$ derivation algorithm, as extinction-derived PM mass concentration is for all particle sizes and $PM_{2.5}$ represents only those particles with diameters smaller than 2.5 μm . A ϕ value of 0.6 was used here, as reported in a past study (Kaku et al., 2018). Equation 1 can thus be rewritten as:

$$C_{m2.5} = \frac{\sigma \times \phi \times 1000}{(\alpha_{scat} \times f_{rh} + \alpha_{abs})} , \quad (3)$$

where $C_{m2.5}$ is CALIOP-derived $PM_{2.5}$ concentration ($\mu g/m^3$).

3 Results and Discussion

3.1 Spatial and temporal patterns of $PM_{2.5}$ from CALIOP and ground-based observations

The yearly variation in $PM_{2.5}$ concentrations from EPA *in situ* instruments across the CONUS was analyzed. A map of the EPA sites with available $PM_{2.5}$ daily data from 2007 to 2018, including our arbitrary boundaries of four geographic regions within the CONUS, provided for context of a subsequent regional analysis, is shown in Fig. 1. Figure 2 shows the $3^\circ \times 3^\circ$ yearly-mean $PM_{2.5}$ concentrations at EPA ground stations from 2007 to 2018, computed from the daily $PM_{2.5}$ measurements. A minimum of fifty observations per $3^\circ \times 3^\circ$ grid box was required in order to be included in the analysis. Consistent spatial patterns of $PM_{2.5}$ generally emerge regardless of year, with higher concentrations in both the eastern and western CONUS (especially California), and lower concentrations across the central CONUS. This pattern is consistent with other studies

(e.g., Hand et al., 2013; Ford et al., 2018; Gantt et al., 2020). In terms of yearly variation, PM_{2.5} in the eastern CONUS is higher early in the study period and decreases in the later years, likely demonstrating the effects of air quality regulation (e.g., Tosca et al., 2017; Hand et al., 2020). For other CONUS regions, it is more difficult to discern any significant yearly variation, as the plots of Fig. 2 suggest little change in PM_{2.5} concentrations over the study period. A detailed trend study is developed in Sect. 3.2.

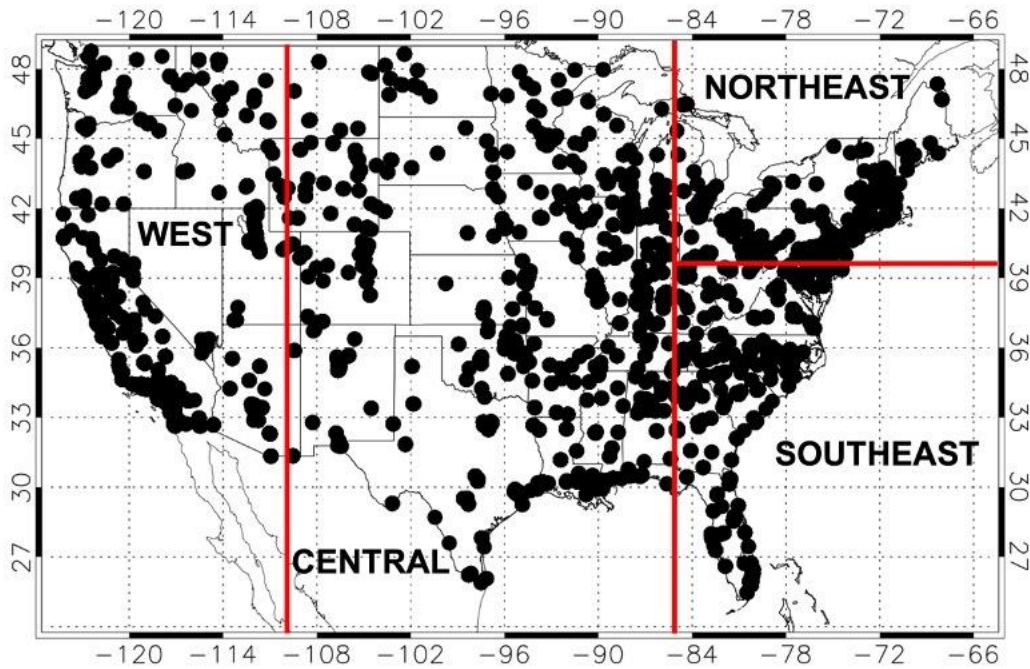


Figure 1. Map of the CONUS showing the locations of U.S. EPA stations that report daily PM_{2.5} concentration observations (Parameter Code: 88101) during the study time period (2007-2018). The red lines delineate the boundaries of four regions: West ($\leq -110^\circ$ longitude), Central ($> -110^\circ$ and $\leq -85^\circ$ longitude), Northeast ($> -85^\circ$ longitude and $\geq 40^\circ$ latitude), and Southeast ($> -85^\circ$ longitude and $< 40^\circ$ latitude).

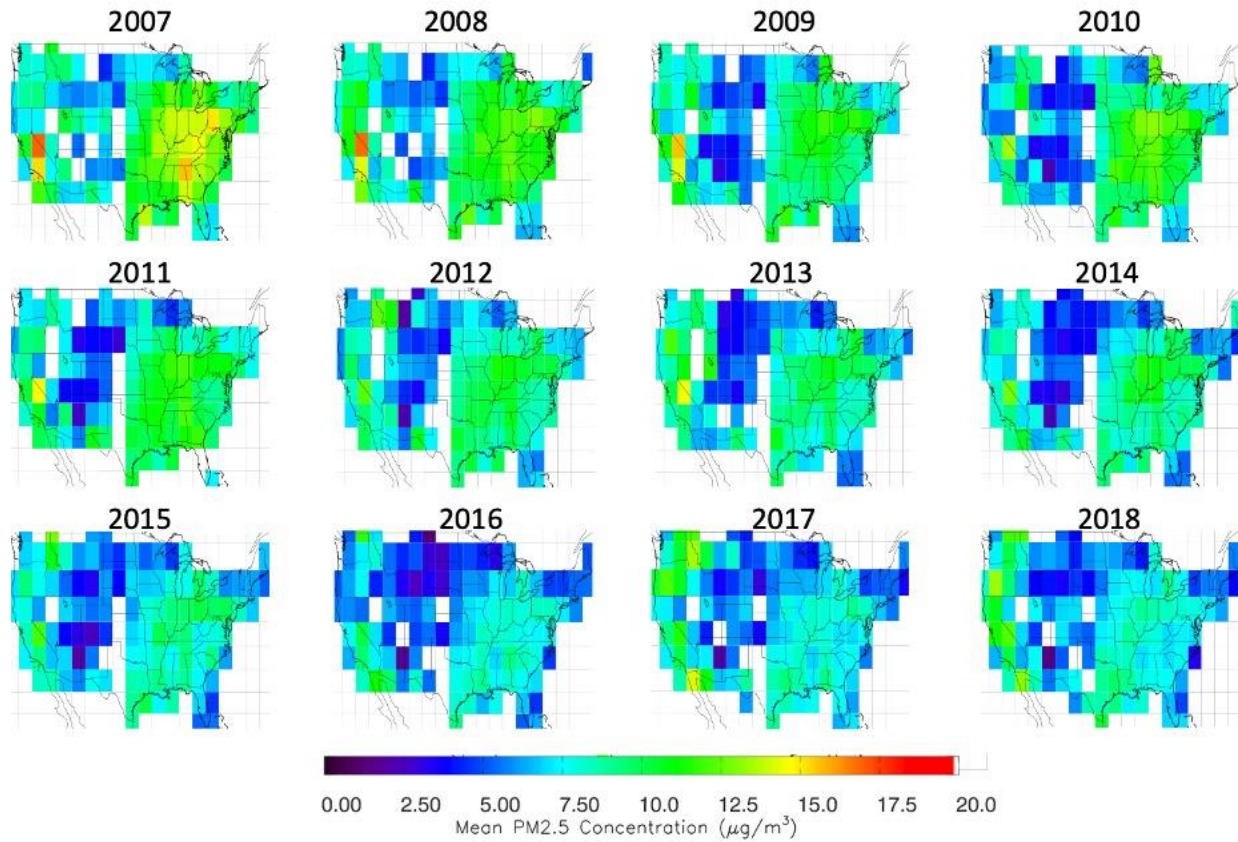


Figure 2. For 2007 to 2018, yearly mean $\text{PM}_{2.5}$ concentrations, computed from daily measurements and gridded at $3^\circ \times 3^\circ$ latitude/longitude resolution, from EPA sites across the CONUS.

We used Eqn. 3 to derive $\text{PM}_{2.5}$ from CALIOP near-surface aerosol extinction (100-1000 m AGL). Using daytime CALIOP measurements (i.e., the near 1330 UTC equator local-time half-orbital granule), $3^\circ \times 3^\circ$ yearly mean $\text{PM}_{2.5}$ concentrations are shown in Fig. 3 for the CONUS over the study period. Note that we require a minimum of fifty points per $3^\circ \times 3^\circ$ grid box, and these yearly means include 100-1000 m aerosol extinction points that are equal to zero from lack of detection sensitivity (e.g., Toth et al., 2018). While there is some noise in the spatial distribution, a broad pattern is observed: higher $\text{PM}_{2.5}$ in the eastern CONUS, lower $\text{PM}_{2.5}$ in the central CONUS, and some areas, like California and Idaho, in the western CONUS with higher $\text{PM}_{2.5}$ concentrations.

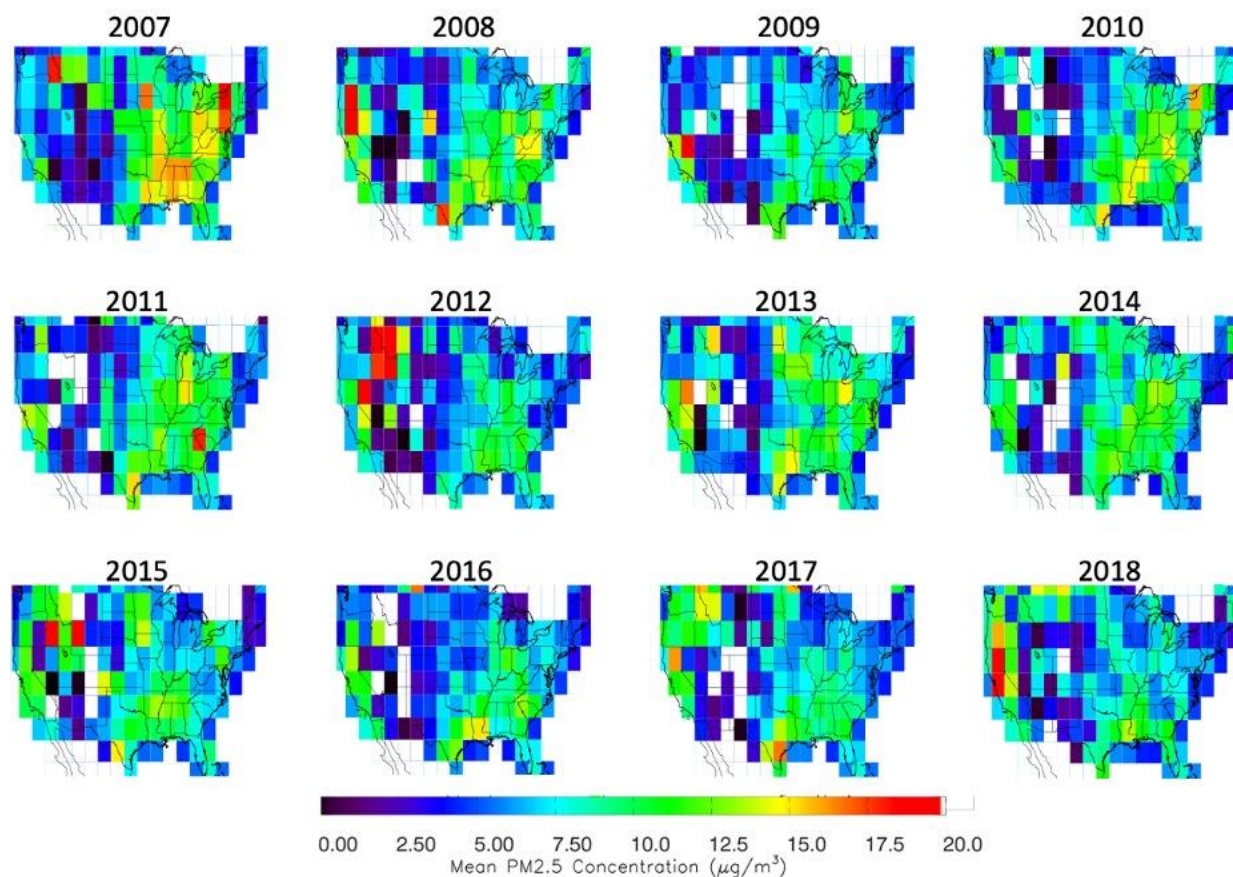


Figure 3. For 2007 to 2018 over the CONUS, $3^\circ \times 3^\circ$ yearly mean $\text{PM}_{2.5}$ concentrations derived from daytime CALIOP near-surface (100-1000 m AGL) aerosol extinction.

Our investigations into these enhanced levels of $\text{PM}_{2.5}$ concentrations in Idaho in 2012, for example, showed they are due to increased wildfire activity in August and September 2012 (e.g., Mallia et al., 2014) that increases the 2012 yearly mean $\text{PM}_{2.5}$ for this area. Also, over the western CONUS, some grid boxes show no data, due in part to the impact of solar contamination in the daytime CALIOP aerosol retrievals. This is exacerbated by the high albedos typically observed in this region compared to other areas in the CONUS (e.g., Houldcroft et al., 2009; Rechid et al., 2009). The lack of data over parts of the western CONUS is also the result of the strict QA protocols, additional screening (cloud-free and dust-free requirements), and data count requirements ($>$ fifty points per grid box) implemented for this study. Note that this data gap

282 (especially during daytime) over parts of the western CONUS with elevated terrain (e.g., Great
 283 Basin and Rockies) is expected and reported in previous studies (e.g., Campbell et al., 2012). This
 284 could have resulted from a number of factors, including solar contamination, high surface albedos,
 285 QA screening procedures, and surface return contamination/topographic effects. Similar to Fig.
 286 3, and with the same data count requirements discussed previously, PM_{2.5} concentrations derived
 287 using nighttime CALIOP aerosol extinction (i.e., the near 0130 UTC equator local-time half-
 288 orbital granule) are shown in Fig. 4. Spatial patterns closely resemble those found during daytime,
 289 but with no data gaps in the western CONUS (now likely due to the lack of solar influence on the
 290 retrievals at night).

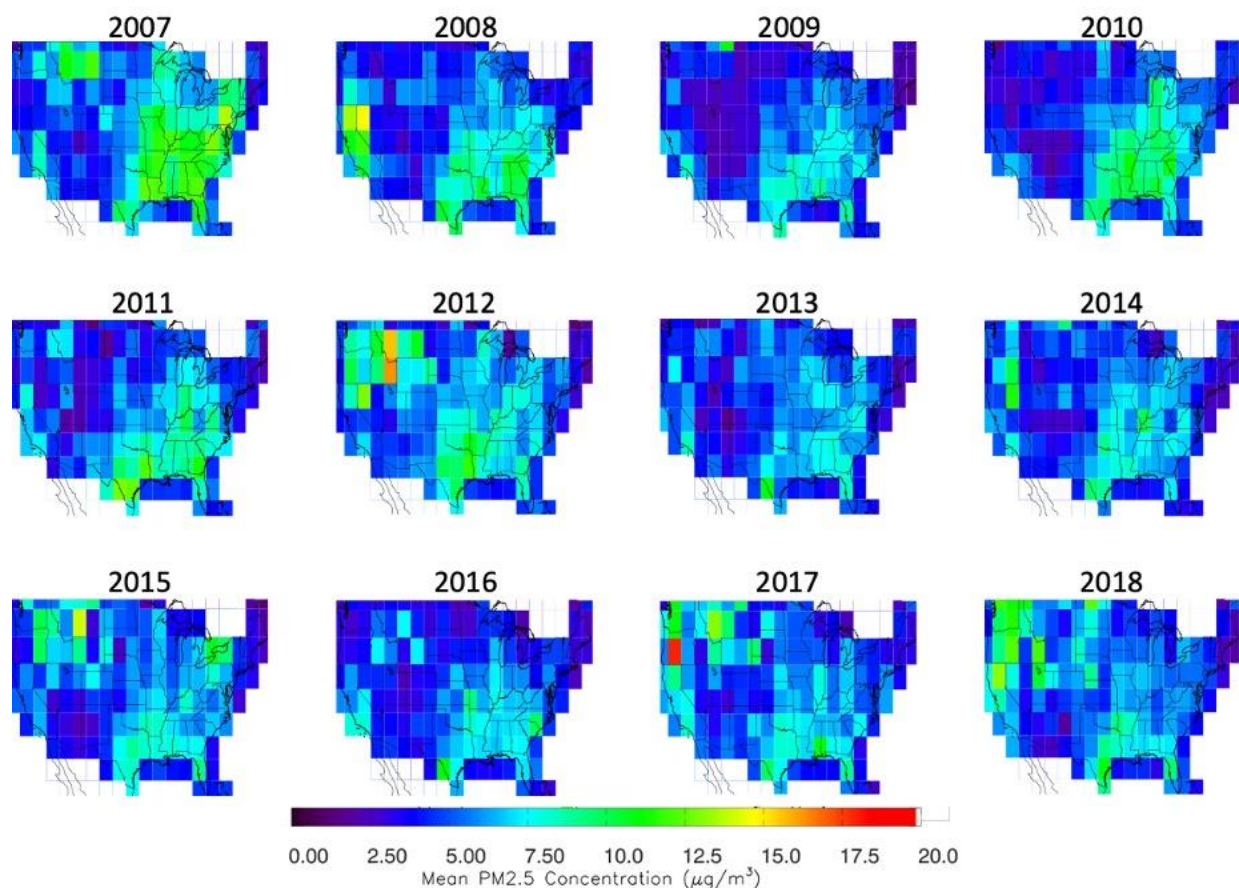


Figure 4. For 2007 to 2018 over the CONUS, 3° x 3° yearly mean PM_{2.5} concentrations derived from nighttime CALIOP near-surface (100-1000 m AGL) aerosol extinction.

Generally, for both daytime and nighttime analyses, a reasonable agreement between the spatial patterns from EPA and those from CALIOP is found (i.e., comparing Fig. 2 with Figs. 3 and 4). Note, however, that the relationship between EPA-based PM_{2.5} and CALIOP-derived PM_{2.5} is explored in a more detailed manner in Sect. 3.2 and 3.3. While some regions and years show better agreement than others, there are instances for which the CALIOP retrieval performs very well. For example, in 2007 over the southeastern CONUS, EPA shows elevated PM_{2.5} concentrations of ~10-15 $\mu\text{g}/\text{m}^3$, with similar values derived from daytime CALIOP aerosol extinction. For context, the corresponding spatial distributions of 3° x 3° yearly-mean CALIOP-based aerosol extinction from which the PM_{2.5} concentrations were derived are shown in the appendix as Supplemental Figs. 1 (daytime data) and 2 (nighttime data).

Lastly for this section, we evaluate the accuracy of the algorithm for the whole study period by spatially (± 100 km) and temporally (same day) collocating CALIOP-derived PM_{2.5} and PM_{2.5} acquired at EPA stations, following the steps mentioned in Toth et al. (2019). In order to reduce the influence of noise and large extinction uncertainties on the analysis, and to reduce the temporal differences between EPA data (i.e., 24-hour measurements) and CALIOP data (i.e., instantaneous measurements), we compute one-year means of PM_{2.5} concentrations from the EPA and CALIOP. One-year means are also chosen to increase the data counts in our analysis due to the ~16 day repeat cycle of the CALIPSO satellite (i.e., ~22 days of observations per year). Consistent with Toth et al. (2019), only EPA stations with one hundred or more collocated EPA/CALIOP data pairs per year were considered for this analysis. Note that we implement a strict data screening process for which we require all 100 m aerosol extinction bins within the 100-1000 m AGL altitude region to be greater than zero for comparison with EPA PM_{2.5} measurements (again, see Toth et al., 2018 for the impact of zero points on CALIOP-based aerosol averages).

314 The results of this exercise are shown in Fig. 5, for separate daytime (Fig. 5a) and nighttime
 315 (Fig. 5b) analyses and a combined daytime/nighttime analysis (Fig. 5c). Each point in the
 316 scatterplots represents a one-year mean $\text{PM}_{2.5}$ concentration from the EPA and CALIOP
 317 throughout the study period, and a Deming regression was fit to the data (Deming, 1943).

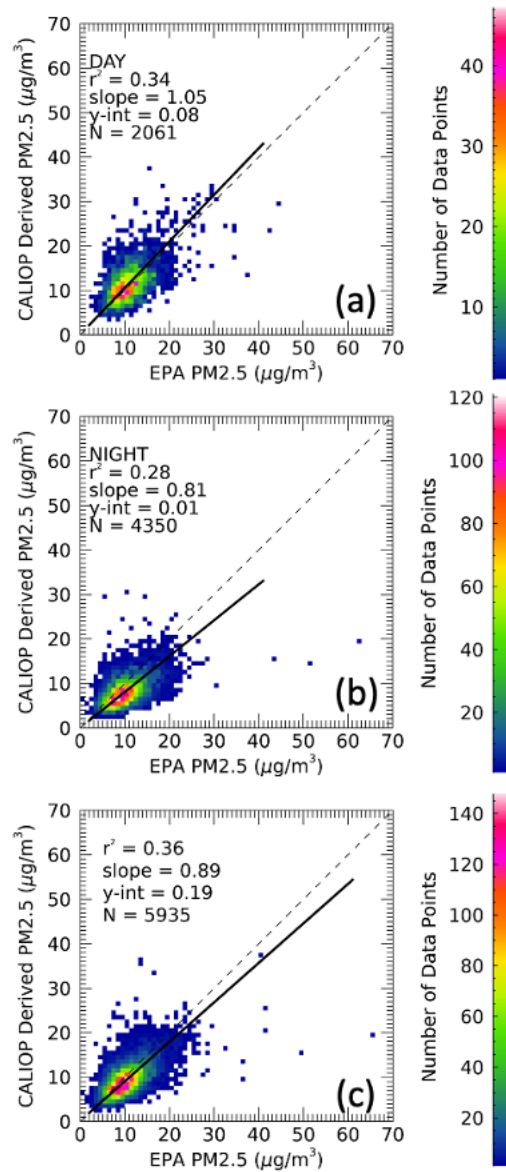


Figure 5. For 2007-2018 over the CONUS, scatterplots of yearly mean $\text{PM}_{2.5}$ concentrations from EPA sites and those derived from collocated near-surface CALIOP aerosol extinction, using (a) daytime, (b) nighttime, and (c) combined daytime and nighttime CALIOP data. Points are color-coded by the number of data points per $1 \mu\text{g}/\text{m}^3$ bin. The dashed and solid lines show the one-to-one lines and Deming regression fits, respectively.

A slightly better agreement between the two datasets is found for daytime ($R^2 = 0.34$; Deming slope = 1.05; RMSE = $3.95 \mu\text{g}/\text{m}^3$; mean bias = $0.62 \mu\text{g}/\text{m}^3$) compared to nighttime ($R^2 = 0.28$; Deming slope = 0.81; RMSE = $3.89 \mu\text{g}/\text{m}^3$; mean bias = $-1.99 \mu\text{g}/\text{m}^3$), with the combined daytime/nighttime analysis yielding the best correlation ($R^2 = 0.36$; Deming slope = 0.89; RMSE = $3.42 \mu\text{g}/\text{m}^3$; mean bias = $-1.00 \mu\text{g}/\text{m}^3$).

Considering our temporal (one day) collocation constraint between CALIOP and EPA data points, these R^2 values are higher than the temporal autocorrelation (i.e., $R^2 \sim 0.25$ for a 24-hour offset) of aerosols reported in Anderson et al. (2003). For our CALIOP/EPA spatial (100 km) collocation constraint, these R^2 values compare reasonably well with the 100 km spatial autocorrelation of synoptic-scale aerosol plumes (i.e., R^2 of ~ 0.30) but are smaller than those using all data (i.e., $R^2 \sim 0.64$), as presented in Anderson et al. (2003). The larger R^2 value for daytime could be due to the relationship of the CALIOP morning/afternoon retrievals to the ground station observed $\text{PM}_{2.5}$ diurnal variability and/or the higher aerosol loadings required for CALIOP aerosol detection during daytime compared to nighttime. It might also be the result of the assumption of the mixed layer height (i.e., 100-100 m AGL) used in this study. These issues are discussed in greater detail in later sections. Figure 5 also shows the data density distributions from 2D histogram computations, as each point is color-coded by the number of points in each $1 \mu\text{g}/\text{m}^3$ bin. For all three analyses (daytime, nighttime, and combined daytime/nighttime), the peak in data counts occurs at $\sim 8\text{-}10 \mu\text{g}/\text{m}^3$.

3.2 Regional analyses

We have further studied regional differences in the mean state and trends of EPA/CALIOP $\text{PM}_{2.5}$ concentrations for four arbitrarily partitioned regions as shown in Fig. 1. The yearly mean

341 PM_{2.5} concentrations from EPA (Fig. 2), daytime CALIOP (Fig. 3), and nighttime CALIOP (Fig.
342 4) over the CONUS were used to compute regional means for the following regions: West ($\leq -$
343 110° longitude), Central ($> -110^\circ$ and $\leq -85^\circ$ longitude), Northeast ($> -85^\circ$ longitude and $\geq 40^\circ$
344 latitude), and Southeast ($> -85^\circ$ longitude and $< 40^\circ$ latitude). We recognize that such large regions
345 will incorporate aerosols from a variety of sources, however a more discrete regional analysis is
346 not the focus of this study.

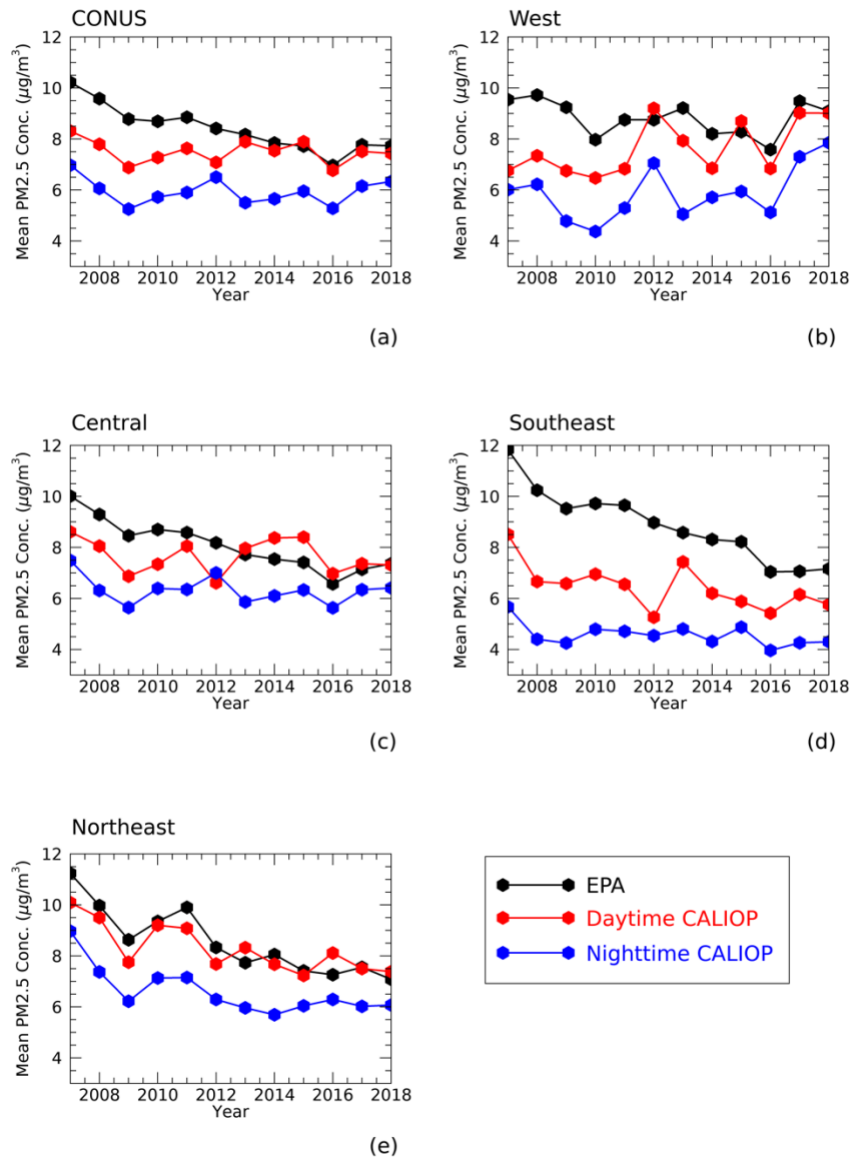


Figure 6. Yearly means of PM_{2.5} concentrations from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP observations for (a) the CONUS and four regions within the CONUS: (b) West, (c) Central, (d) Southeast, and (e) Northeast.

Figure 6 shows the yearly mean regional PM_{2.5} concentrations from EPA (in black), daytime CALIOP (in red), and nighttime CALIOP (in blue) observations for the entire CONUS (Fig. 6a) and four regions: West (Fig. 6b), Central (Fig. 6c), Southeast (Fig. 6d), and Northeast (Fig. 6e). For the CONUS, the yearly mean PM_{2.5} concentrations for EPA more closely match

daytime CALIOP PM_{2.5} than nighttime CALIOP PM_{2.5}. Also, nighttime CALIOP PM_{2.5} are consistently smaller than EPA and daytime CALIOP PM_{2.5} for the CONUS throughout the study period, which may plausibly be due to a low bias in near surface CALIOP aerosol extinction at nighttime, or from the diurnal differences in PM_{2.5} concentrations. Both of these patterns for the CONUS are evident for each of the four regions as well.

The region with the largest yearly mean PM_{2.5} differences between the datasets is the Southeast (e.g., 2007, 2008, and 2012). The differences may be attributed to sampling differences between observations from surface stations and CALIOP. It is also possible the differences are caused by retrieval related biases. To further investigate the differences, we compared pair wise the CALIOP-derived yearly mean PM_{2.5} concentrations against those of collocated EPA station data, using methods discussed in Toth et al. (2019) and later in this section, for each of the four regions mentioned above (not shown). CALIOP-EPA PM_{2.5} mean biases on the order of a few $\mu\text{g}/\text{m}^3$ are found for each region. For example, during daytime, PM_{2.5} mean biases range from \sim 0.5 $\mu\text{g}/\text{m}^3$ (Southeast) to \sim 2.3 $\mu\text{g}/\text{m}^3$ (West). During nighttime, mean bias values range from \sim 3.5 $\mu\text{g}/\text{m}^3$ (Southeast) to \sim 1.6 $\mu\text{g}/\text{m}^3$ (Central). These differences are smaller than the differences shown in Fig. 5d, suggesting some of the differences of Fig. 5d are from sampling related biases.

We have also conducted a seasonal analysis of the EPA- and CALIOP- based PM_{2.5} concentrations for the CONUS and the four regions within the CONUS as previously discussed. Monthly mean PM_{2.5} concentrations from each dataset for the 2007-2018 time period are shown in Fig. 7. For the CONUS (Fig. 7a), EPA PM_{2.5} levels are largest in August (\sim 10 $\mu\text{g}/\text{m}^3$), with elevated levels in December and January and the lowest levels during the spring and fall months. Daytime and nighttime CALIOP-based PM_{2.5} concentrations agree well for May through

September ($5\text{--}6\text{ }\mu\text{g}/\text{m}^3$). As in the EPA time series, peak daytime $\text{PM}_{2.5}$ levels in the CALIOP data are found during wintertime, while the nighttime CALIOP peak occurs in August.

The seasonal patterns observed for the entire CONUS are generally consistent regionally, but $\text{PM}_{2.5}$ concentration peaks at higher levels for some regions. For example, in the West (Fig. 7b), the maximum monthly mean $\text{PM}_{2.5}$ concentrations is $\sim 12\text{ }\mu\text{g}/\text{m}^3$ in August (e.g., possibly due to the fire season; Mallia et al., 2014) and during December and January. For the Southeast (Fig. 7d), the daytime CALIOP $\text{PM}_{2.5}$ peak is $\sim 13\text{ }\mu\text{g}/\text{m}^3$ (occurring in February), and the Northeast region (Fig. 7e) exhibits the most agreement between all three datasets (with $\text{PM}_{2.5}$ levels peaking during July).

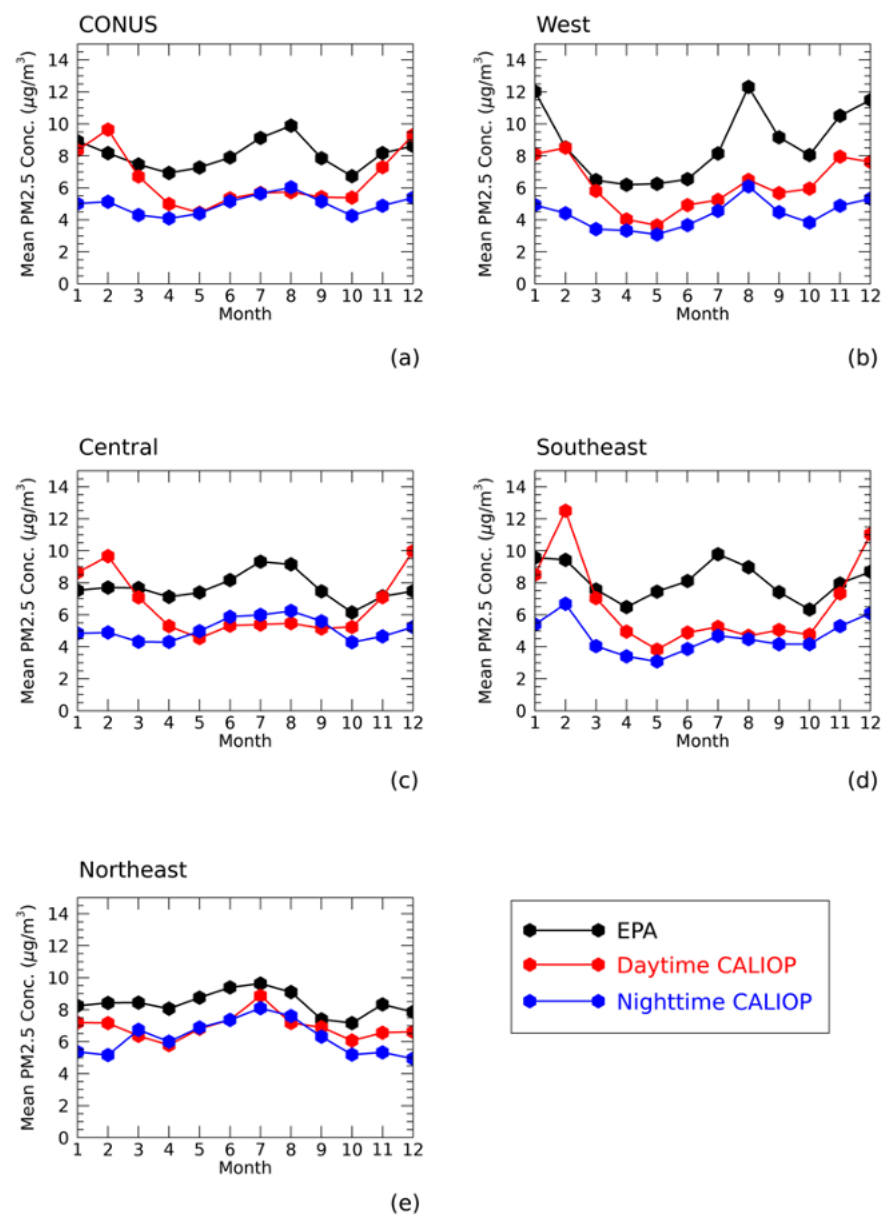


Figure 7. Monthly means of PM_{2.5} concentrations from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP observations for (a) the CONUS and four regions within the CONUS: (b)West, (c) Central, (d) Southeast, and (e) Northeast.

3.3 Long-term variations of PM_{2.5} from CALIOP and ground-based observations

In this section, the twelve-year mean and trends of EPA and CALIOP-based PM_{2.5} concentrations over the CONUS are explored. Figure 8a shows 3° x 3° mean PM_{2.5} at EPA stations for the twelve-year study period, and the spatial distribution follows the same pattern as discussed earlier in the paper: higher concentrations in the eastern/western CONUS and lower in the central CONUS. The corresponding twelve-year 3° x 3° mean PM_{2.5} concentrations derived from CALIOP are shown in Fig. 8b (daytime CALIOP) and Fig. 8c (nighttime CALIOP).

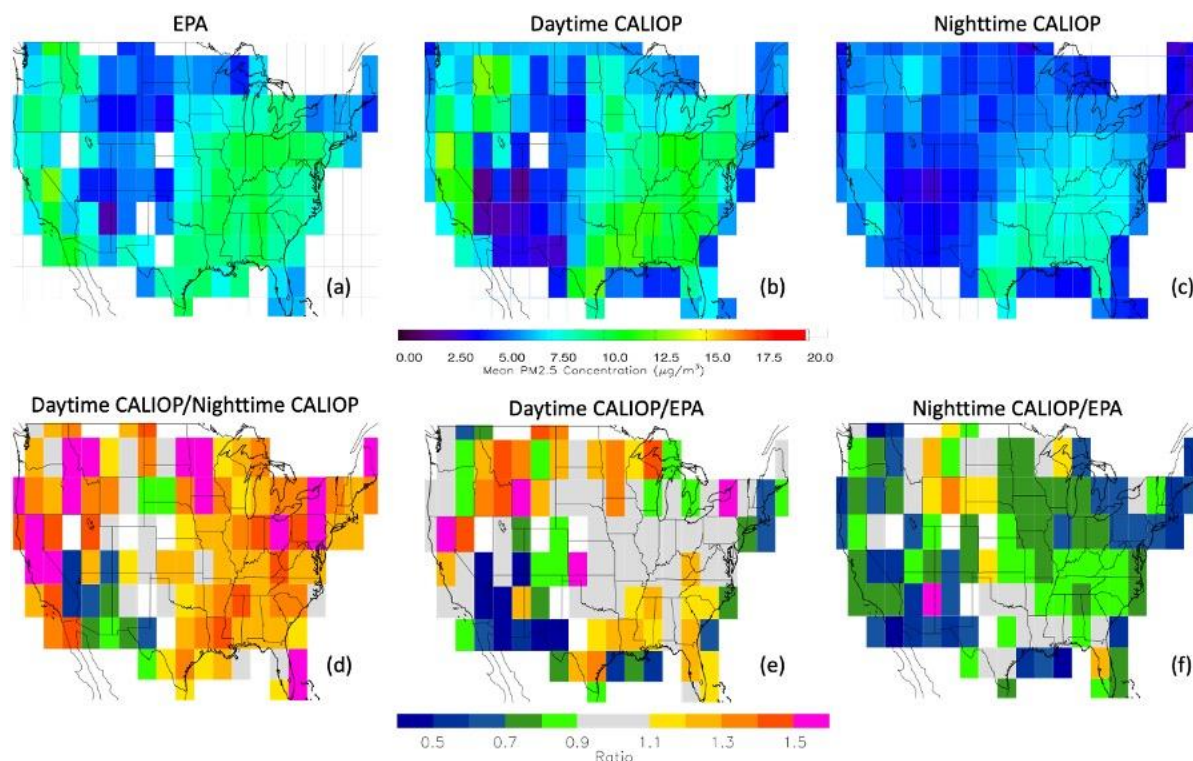


Figure 8. Twelve-year (2007-2018) mean PM_{2.5} concentrations (a) at EPA stations and those derived from (b) daytime, and (c) nighttime, CALIOP near-surface aerosol extinction (gridded at 3° x 3° latitude/longitude resolution). Also shown are the corresponding ratios of (d) daytime to nighttime CALIOP-derived PM_{2.5}, (e) daytime CALIOP PM_{2.5} to EPA PM_{2.5}, and (f) nighttime CALIOP PM_{2.5} to EPA PM_{2.5}, computed for only those grid boxes with available data for each of the analyses in Fig. 8a-c.

Here we require a minimum of 600 points per 3° x 3° grid box (i.e., 50 points per year for 12-year period), and these means include 100-1000 m aerosol extinction equal to zero points. While

daytime CALIOP $\text{PM}_{2.5}$ are larger than nighttime CALIOP $\text{PM}_{2.5}$, both show similar spatial patterns that match well with that from the EPA. The difference in daytime and nighttime CALIOP-derived $\text{PM}_{2.5}$ concentrations is consistent with our past study using two years of CALIOP data (Toth et al., 2019), and is a result of the larger mean near-surface aerosol extinction found over the CONUS during daytime compared to nighttime. While these larger aerosol extinction coefficients may be due to increased aerosol amount near the surface (e.g., from elevated anthropogenic emissions during daytime), solar contamination of the daytime CALIOP aerosol extinction retrieval may also be a factor.

We also computed the corresponding ratios of daytime CALOP-derived $\text{PM}_{2.5}$ to nighttime CALIOP-derived $\text{PM}_{2.5}$ (Fig. 8d), daytime CALIOP-derived $\text{PM}_{2.5}$ to EPA-based $\text{PM}_{2.5}$ (Fig. 8e), and nighttime CALIOP-derived $\text{PM}_{2.5}$ to EPA-based $\text{PM}_{2.5}$ (Fig. 8f), for those grid boxes with available data for each of the analyses (i.e., Fig. 8a-c). Most areas of the CONUS show daytime CALIOP-derived $\text{PM}_{2.5}$ concentration estimates larger than those derived from nighttime CALIOP data, with several grid boxes exhibiting ratios greater than 1.5 (Fig. 8d). The exception is for a few grid boxes over the Great Plains and the southwest CONUS, for which daytime CALIOP-derived $\text{PM}_{2.5}$ values are smaller than those of nighttime CALIOP-derived $\text{PM}_{2.5}$. In terms of the relationship between daytime CALIOP $\text{PM}_{2.5}$ and EPA $\text{PM}_{2.5}$, many grid boxes with ratios ~ 1 are found throughout the CONUS (Fig. 8e). Southwest CONUS generally exhibits smaller $\text{PM}_{2.5}$ concentrations from daytime CALIOP data than EPA observations (likely due to the omission of dust layers from the analysis), otherwise most areas are fairly scattered (i.e., similar number of grid boxes with ratios less than and greater than 1) with no clear regional patterns. However, this is not true for the relationship between nighttime CALIOP $\text{PM}_{2.5}$ and EPA $\text{PM}_{2.5}$, for which most

grid boxes exhibit smaller nighttime CALIOP-derived PM_{2.5} concentrations than those from EPA (Fig. 8f).

Note that the diurnal variation of PM_{2.5} pollution is dependent on region and season (due in part to meteorology and boundary layer dynamics, e.g., Sun et al., 2013; Chu et al., 2013) and thus may differ even on a site-by-site basis. Furthermore, the CALIOP datasets provide instantaneous estimates of PM_{2.5} concentrations at one time for each daytime and nighttime overpass for a given location. Thus, it is challenging to characterize the expected daytime-nighttime differences in CALIOP-based PM_{2.5} concentrations broadly over the CONUS. The relationship of the PM_{2.5} concentrations estimated at the daytime and nighttime CALIOP overpass times to the EPA-observed PM_{2.5} diurnal variation will be explored in a future study.

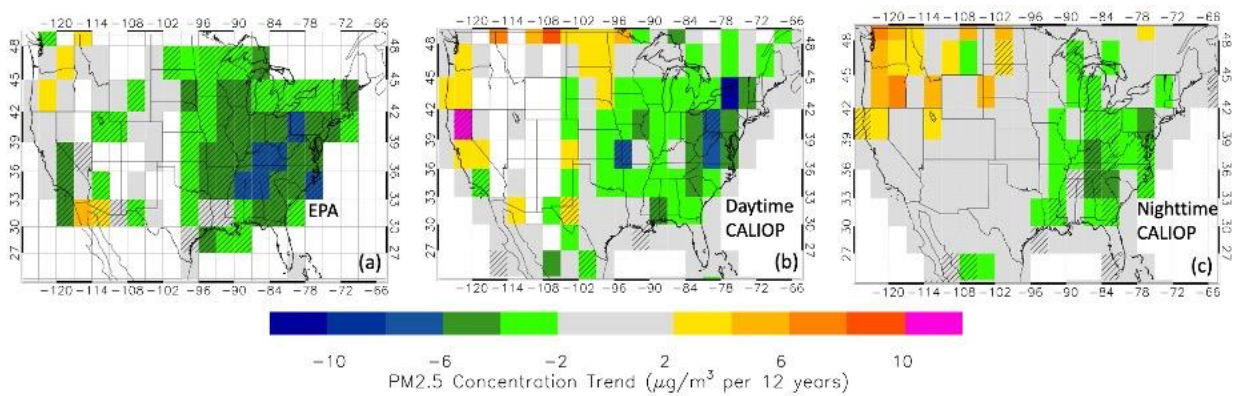


Figure 9. Twelve-year (2007-2018) PM_{2.5} concentration trends (a) at EPA stations and those computed from CALIOP measurements for (b) daytime, and (c) nighttime, conditions (gridded at 3° x 3° latitude/longitude resolution). Hatched grid boxes indicate trends that are significant at the 95% confidence interval, calculated using the Mann-Kendall Test.

Figure 9 shows the spatial variability of the 3° x 3° twelve-year trends of PM_{2.5} at EPA stations (Fig. 9a). These trends are computed as the slope from simple linear regression, with the requirement that for a particular grid box every year in the twelve-year period must be represented. If this requirement is not met, no trend is reported. Figure 9a reveals that from 2007 to 2018, most

EPA-based PM_{2.5} concentrations in the eastern CONUS are negative (about -4 to -6 µg/m³ per 12 years), indicating an improvement in air quality over the study period. Over the central CONUS, most EPA-reported PM_{2.5} concentrations exhibit slightly negative/near-zero PM_{2.5} trends (i.e., little change in air quality with time). This is generally true for the western CONUS as well, with the exception of some grid boxes in the Pacific Northwest and Southwest that show positive PM_{2.5} trends (about 2-4 µg/m³ per 12 years), implying a worsening of air quality from 2007 to 2018 (e.g., due to wildfire activity; McClure and Jaffe, 2018).

Trends for CALIOP-derived PM_{2.5} are shown in Fig. 9b (daytime CALIOP) and Fig. 9c (nighttime CALIOP), computed in the same manner as the EPA analysis, and are only reported if each 3° x 3° grid box included each year in the twelve-year period represented. While the daytime analysis reveals negative PM_{2.5} trends for the eastern CONUS, fewer trends are reported over the western CONUS. The reason for this is the strict data availability requirement implemented for the trend computation, which is a difficult criterion to meet, likely because of terrain effects for the western CONUS and the dust-free requirements. For those grid boxes in the western CONUS with daytime CALIOP trends available, most are positive, especially in California, the Pacific Northwest, and Northern Great Plains. For nighttime (Fig. 9c), the western CONUS reports a far greater number of trend estimates compared to the daytime analysis. While most of these are near zero, positive trends in PM_{2.5} are found over the northwestern CONUS and northern Great Plains. Negative PM_{2.5} trends are found for the eastern CONUS, a result that agrees with the trends computed from both daytime CALIOP data and EPA data. It is noteworthy that the trend patterns discussed here for both EPA and CALIOP are consistent with those values computed using the yearly mean/regional PM_{2.5} concentrations of Fig. 6, as reported in Table 1.

	PM_{2.5} Concentration Trend (µg/m³per 12 years)		
Region	EPA	Daytime CALIOP	Nighttime CALIOP
CONUS	-2.71	-0.39	-0.28
West	-0.77	2.12	1.66
Central	-3.01	-0.51	-0.58
Southeast	-4.58	-1.98	-0.79
Northeast	-3.9	-2.46	-2.25

Table 1. Twelve-year (2007-2018) trends in PM_{2.5} concentrations (µg/m³) at EPA stations and derived from daytime and nighttime CALIOP observations for the CONUS and four regions within the CONUS: West, Central, Southeast, and Northeast. The trends were computed using the yearly mean PM_{2.5} concentrations shown in Fig. 5.

The grid boxes of Figs. 9a-9c with a hatched pattern show the locations of statistically significant PM_{2.5} trends from EPA, daytime CALIOP, and nighttime CALIOP, data. The Mann-Kendall test was used (Mann, 1945; Kendall, 1975; Yue et al., 2002; Toth et al., 2016), and we report trends that are significant at the 95% confidence interval. This analysis reveals that the portion of PM_{2.5} trends from EPA data that are statistically significant is considerably greater than that from CALIOP data. This is possibly due to the uncertainties in CALIOP aerosol extinction (e.g., Young et al., 2013; 2018) from which the PM_{2.5} concentrations are derived.

To further compare the PM_{2.5} trends between EPA and CALIOP, we focus only on those grid boxes that exhibit statistically significant PM_{2.5} trends from either daytime (Fig. 9b) or nighttime (Fig. 9c) CALIOP data. The results of this analysis are shown in Table 2. For the daytime CALIOP analysis, most CALIOP and EPA PM_{2.5} trends compare well with one another, are negative, and also exhibit statistically significant EPA trends. For the daytime CALIOP analysis, most CALIOP PM_{2.5} trends are negative and compare well with those from EPA data. However, there are two grid boxes that show positive PM_{2.5} trends for both CALIOP and EPA, and

one of which (i.e., the -121.5° longitude midpoint, 40.5° latitude midpoint grid box; northern California) shows a particularly large difference in magnitude (i.e., $\sim 10 \mu\text{g}/\text{m}^3$ for CALIOP but $\sim 0 \mu\text{g}/\text{m}^3$ for EPA). This is likely due to differences in sampling between the CALIOP and EPA datasets, especially in areas with highly variable terrain (e.g., the valleys, mountains, and forests of northern California). For the nighttime CALIOP analysis, there are a greater number of grid boxes with statistically significant trends, most of which are found in the eastern CONUS, are negative, and agree well with EPA $\text{PM}_{2.5}$ trends. Also, note that most of the EPA trends associated with both daytime and nighttime statistically significant CALIOP trends reported in Table 1 are also statistically significant at the 95% CI.

DAYTIME CALIOP ANALYSIS				
		Trend ($\mu\text{g}/\text{m}^3$ per 12 years)		
Longitude Midpoint (deg.)	Latitude Midpoint (deg.)	CALIOP	EPA	Significant (95% CI) EPA Trend?
-121.5	40.5	10.55	0.03	No
-118.5	34.5	-3.41	-5.28	Yes
-103.5	31.5	3.82	0.79	No
-94.5	37.5	-6.09	-4.18	Yes
-91.5	28.5	-1.14	-2.23	Yes
-88.5	31.5	-4.30	-3.90	Yes
-79.5	40.5	-7.67	-6.16	Yes
NIGHTTIME CALIOP ANALYSIS				
		Trend ($\mu\text{g}/\text{m}^3$ per 12 years)		
Longitude Midpoint (deg.)	Latitude Midpoint (deg.)	CALIOP	EPA	Significant (95% CI) EPA Trend?
-124.5	40.5	2.47	0.01	No
-100.5	46.5	3.47	-2.32	Yes
-94.5	28.5	-1.99	-4.83	Yes
-91.5	31.5	-4.00	-1.84	Yes
-88.5	34.5	-1.94	-6.08	Yes
-88.5	46.5	-3.85	-2.81	Yes
-85.5	34.5	-5.28	-6.48	Yes
-85.5	37.5	-4.67	-6.08	Yes
-85.5	40.5	-2.82	-5.96	Yes
-85.5	43.5	-2.18	-3.26	Yes
-82.5	31.5	-3.66	-4.56	Yes
-82.5	37.5	-3.50	-6.59	Yes
-82.5	40.5	-2.10	-5.55	Yes
-76.5	34.5	-3.89	-6.37	Yes
-76.5	37.5	-3.60	-4.67	Yes

Table 2. The PM_{2.5} trends and their locations for those 3° x 3° grid boxes with statistically significant (95% CI) daytime or nighttime CALIOP PM_{2.5} trends, and the corresponding EPA PM_{2.5} trends, as determined from Fig. 6. The corresponding EPA PM_{2.5} trends that are statistically significant (95% CI) are also indicated.

479

480

Region	PM _{2.5} Concentration Trend (ug/m ³ per 12 years)											
	EPA				Daytime CALIOP				Nighttime CALIOP			
	DJF	MAM	JJA	SON	DJF	MAM	JJA	SON	DJF	MAM	JJA	SON
CONUS	-3.25	-2.86	-2.12	-2.54	-0.88	-0.51	-0.52	0.4	-0.1	-0.77	-0.4	-0.13
West	-2.99	-1.71	2.56	-0.9	-2.75	-0.22	2.53	1.54	0.03	-0.27	2.04	1.04
Central	-3.57	-2.99	-2.41	-3.03	-0.7	-0.89	-0.45	0.01	-0.12	-1.06	-0.46	-0.43
Southeast	-3.36	-3.87	-6.67	-3.87	-0.16	1.27	-3.16	0.39	0.06	0.03	-1.26	-0.23
Northeast	-2.36	-3.84	-6.87	-2.59	-0.81	-2.6	-2.5	-0.29	-0.51	-1.75	-4.35	-1.12

Table 3. Twelve-year (2007-2018) trends in PM_{2.5} concentrations (µg/m³) at EPA stations and derived from daytime and nighttime CALIOP observations for the CONUS and four regions within the CONUS for each season: December, January, February (DJF), March, April, May (MAM), June, July, August (JJA), and September, October, November (SON). The trends were computed using the seasonal mean PM_{2.5} concentrations shown in Supplemental Figs. 3-6.

The analyses of this dataset have revealed regional differences in PM_{2.5} trends throughout the CONUS (e.g., Fig. 9 and Table 1). To investigate the reason for these differences, we explore the seasonality of the twelve-year (2007-2018) temporal variations of PM_{2.5} observed at EPA sites and derived from daytime and nighttime CALIOP measurements. Trends were computed from simple linear regression using the seasonal/regional mean PM_{2.5} concentrations provided in the appendix as Supplemental Figs. 3-6. The results of this analysis are shown in Table 3, segmented by region and four seasons: December, January, February (DJF), March, April, May (MAM), June, July, August (JJA), and September, October, November (SON). The majority of regions and seasons exhibit negative PM_{2.5} trends on the order of a few µg/m³ to as much as ~ -7 µg/m³ (JJA

for Southeast and Northeast). The region that stands out is the West, for which several positive $PM_{2.5}$ trends are found. These positive trends mostly occur during the JJA and SON seasons, but negative trends in $PM_{2.5}$ are found in the West for other seasons. This pattern reveals two important points about the seasonality of $PM_{2.5}$ concentrations for the western CONUS compared to the eastern CONUS. For one, during the more active fire seasons in the West (JJA and SON), $PM_{2.5}$ trends are positive, demonstrating the impact of wildfires on $PM_{2.5}$ concentrations. This likely explains the differences in $PM_{2.5}$ trends for the CONUS as a whole and regionally (e.g., Fig. 9 and Table 1). Second, during the other two seasons (DJF and MAM), $PM_{2.5}$ trends in the West are negative, and thus are in better agreement with the trends exhibited over eastern CONUS. This likely indicates that air quality regulations are functioning in a similar manner for both the eastern and western CONUS.

3.4 Mixed layer height sensitivity studies

The sensitivity of the parameters in Eqn. 3 to the CALIOP retrieved $PM_{2.5}$ concentrations has been explored in Toth et al. (2019), and thus we do not repeat most of the exercises implemented in that paper. Still, consistent with Toth et al. (2019), the mixed layer for this paper is assumed to be 100-1000 m AGL, and thus only CALIOP aerosol extinction in this altitude range is used to derive $PM_{2.5}$ concentrations. The lowest 100 m was not considered due to the possibility of surface contamination in the aerosol retrieval (e.g., Kim et al., 2017). Since only two years of data were used in Toth et al. (2019), we have further tested these analyses using twelve years of the CALIOP data record. Also, we have explored this sensitivity by both considering and excluding the layer nearest the surface (i.e., 0-100 m AGL), as well as extending the surface-height analysis up to 2 km AGL. Also, here we require each 100 m aerosol extinction bin to be greater than zero when varying the height of the assumed mixed layer.

515 The results of the sensitivity study including the 0-100 m AGL layer are shown in Table 4.
516 In general, varying the height of the mixed layer shows no major changes to the EPA/CALIOP
517 PM_{2.5} relationship for the daytime analysis, with R^2 values around 0.2-0.3 for most layers.
518 However, the smallest R^2 value is found for the 0-100 m layer during nighttime ($R^2 = 0.05$),
519 corresponding to the largest mean bias (CALIOP-EPA) of about 6.7 $\mu\text{g}/\text{m}^3$. This is consistent with
520 Toth et al. (2019), yet the poor performance of the CALIOP PM_{2.5} retrieval, including whether or
521 not surface return contamination is a factor here, requires further investigation. Contrary to the
522 daytime analysis, the nighttime R^2 values monotonically increase as the mixed layer thickness
523 increases.

Analysis (Day/Night)			
Mixed Layer (m)	R ²	Deming Slope	Mean Bias (CALIOP - EPA; $\mu\text{g}/\text{m}^3$)
0-100	0.23/0.05	1.09/0.56	-2.79/-6.72
0-200	0.27/0.19	1.19/0.72	-0.13/-3.99
0-300	0.29/0.23	1.19/0.77	0.41/-3.23
0-400	0.31/0.25	1.15/0.77	0.40/-2.87
0-500	0.31/0.27	1.12/0.76	0.38/-2.66
0-600	0.31/0.29	1.10/0.74	0.29/-2.56
0-700	0.31/0.29	1.08/0.72	0.30/-2.55
0-800	0.28/0.29	1.08/0.71	0.24/-2.50
0-900	0.25/0.31	1.08/0.74	0.17/-2.45
0-1000	0.26/0.32	1.08/0.75	0.16/-2.45
0-1100	0.30/0.32	1.07/0.77	0.15/-2.43
0-1200	0.29/0.34	1.07/0.75	0.11/-2.45
0-1300	0.26/0.35	1.05/0.74	0.16/-2.46
0-1400	0.23/0.36	1.01/0.74	0.20/-2.43
0-1500	0.24/0.37	1.01/0.75	0.17/-2.45
0-1600	0.25/0.39	1.02/0.76	0.19/-2.41
0-1700	0.26/0.40	1.03/0.76	0.16/-2.41
0-1800	0.27/0.41	1.07/0.79	0.18/-2.28
0-1900	0.23/0.43	1.06/0.80	0.02/-2.28
0-2000	0.22/0.45	1.03/0.78	-0.18/-2.31

Table 4. Results of a sensitivity study varying the height of the assumed mixed layer, including R², slope computed from Deming regression analysis, and mean bias (CALIOP – EPA; $\mu\text{g}/\text{m}^3$). This analysis includes the layer nearest the surface (0-100 m).

Table 5 shows the results of the sensitivity study excluding the 0-100 m AGL layer. There is little variability in the EPA/CALIOP PM_{2.5} R² values when adjusting the assumed height of the mixed layer for both the daytime and nighttime analyses. However, there is a near monotonic decrease in the daytime CALIOP-EPA bias as the thickness of the assumed mixed layer, yet this pattern not evident in the nighttime biases. We note the results shown in Tables 4 and 5 could be less indicative of the actual aerosol vertical distribution near the surface but more due to the fact

530 that we apply one mixed layer height value for the entire CONUS, in addition to the other reasons
531 discussed below.

Analysis (Day/Night)			
Mixed Layer (m)	R²	Deming Slope	Mean Bias (CALIOP - EPA; $\mu\text{g}/\text{m}^3$)
100-200	0.23/0.22	1.28/0.94	2.37/-2.67
100-300	0.25/0.24	1.25/0.93	2.16/-2.18
100-400	0.27/0.25	1.19/0.91	1.75/-2.08
100-500	0.28/0.26	1.15/0.88	1.53/-2.00
100-600	0.32/0.27	1.10/0.86	1.24/-1.99
100-700	0.37/0.28	1.04/0.84	1.04/-1.96
100-800	0.35/0.28	1.03/0.82	0.92/-1.96
100-900	0.35/0.29	1.04/0.81	0.71/-1.97
100-1000	0.34/0.28	1.05/0.81	0.62/-1.99
100-1100	0.33/0.30	1.04/0.83	0.58/-2.01
100-1200	0.35/0.31	1.06/0.82	0.46/-2.05
100-1300	0.34/0.33	1.05/0.82	0.47/-2.08
100-1400	0.33/0.34	1.04/0.81	0.46/-2.10
100-1500	0.32/0.35	1.05/0.81	0.43/-2.12
100-1600	0.32/0.35	1.04/0.80	0.37/-2.16
100-1700	0.31/0.35	1.03/0.78	0.35/-2.21
100-1800	0.31/0.34	1.07/0.76	0.31/-2.25
100-1900	0.33/0.34	1.04/0.73	0.06/-2.30
100-2000	0.31/0.32	1.03/0.72	0.05/-2.28

Table 5. Results of a sensitivity study varying the height of the assumed mixed layer, including R², slope computed from Deming regression analysis, and mean bias (CALIOP – EPA; $\mu\text{g}/\text{m}^3$). This analysis excludes the layer nearest the surface (0-100 m).

532 We recognize that conducting these sensitivity studies of mixed layer height over the
533 CONUS as a whole is a simplification, as the mixed layer/PBL height varies seasonally and
534 regionally and is dependent upon the local meteorology (e.g., Seidel et al., 2012; Zhang et al.,
535 2020). A more in depth seasonal and regional study of this parameter and its impact on our
536 CALIOP-derived PM_{2.5} estimates is needed. However, the findings of our mixed layer height

sensitivity studies could also be due to aerosol spatial autocorrelation lengths and the EPA/CALIOP collocation limits applied in this study.

For example, by using the data curtain provided by lidars to derive $PM_{2.5}$ concentrations, if the surface term (i.e., 0-100 m) alone is utilized unrestrained, the limited spatial correlation lengths at the surface can negatively impact the surface station/lidar $PM_{2.5}$ relationship. This is especially likely over the complex terrain of the western CONUS, for which shorter EPA surface station - observed $PM_{2.5}$ spatial correlation lengths are found compared to the entire CONUS (Toth et al., 2019). Rather, we might be better served to average extinction to higher altitudes (e.g., 1-2 km AGL), whereby spatial correlation lengths are possibly longer at such heights and thus would characterize near-surface aerosol extinction in a more stable manner, rendering a better comparison and $PM_{2.5}$ product overall.

We note that it is also entirely plausible that the 100 m bin closest to the surface alone is fully accurate, and that the collocation and subsequent analysis applied here are unrepresentative of the actual performance of the 0-100 m bin alone (i.e., if these points were directly collocated, it is possible that the relationship is the inverse of what is shown for this analysis). As Omar et al. (2013) report, however, collocation of ground observations with the CALIOP curtain is challenging. The reader is thus encouraged to consider the context outlined above in interpreting the results of our sensitivity studies and in deciding how exactly to apply either methodology.

3.5 Surface-to-column aerosol representativeness analysis

One assumption under the traditional AOT and $PM_{2.5}$ analyses is that column-integrated AOT can be used to represent near-surface aerosol concentration (e.g., Hoff and Christopher, 2009). Toth et al. (2014) investigated surface-to-column aerosol representativeness using 2 years of CALIOP data and reported a large spatial variation in the fraction of 100-1000 m AOT to the

total column AOT. Given the long record of CALIPSO data, it is intriguing to study the temporal variation of the surface-to-column aerosol representativeness as measured by CALIOP. While it is ideal to use CALIOP-derived PBL height estimates here rather than the assumed 100-1000 m AGL layer, this task is challenging, and thus we believe such a topic is best suited for its own paper.

Using twelve years of CALIOP cloud-free and dust-free aerosol profile data (i.e., no clouds or dust aerosols identified in the entire atmospheric column), we study the mean state of CALIOP-observed surface-to-column aerosol representativeness over the CONUS. Figure 10a shows the $3^\circ \times 3^\circ$ mean fraction of daytime CALIOP 100-1000 m AOT to the total column AOT (%) over the CONUS for the study period. The same data count requirements as Figs. 8b and 8c are implemented for Fig. 10, but with the addition of a near-surface aerosol metric (i.e., the presence of aerosol is required for at least one bin in the 100-1000 m altitude region). We note the lack of daytime data over the western CONUS, due to issues with CALIOP data as discussed earlier in the paper. Over most of the CONUS, the daytime contribution of the AOT below 1 km to the total column AOT is in the neighborhood of 50% to 60% (i.e., roughly consistent with Reid et al., 2017), with the largest contributions occurring in the western CONUS (e.g., California, Oregon, and Washington).

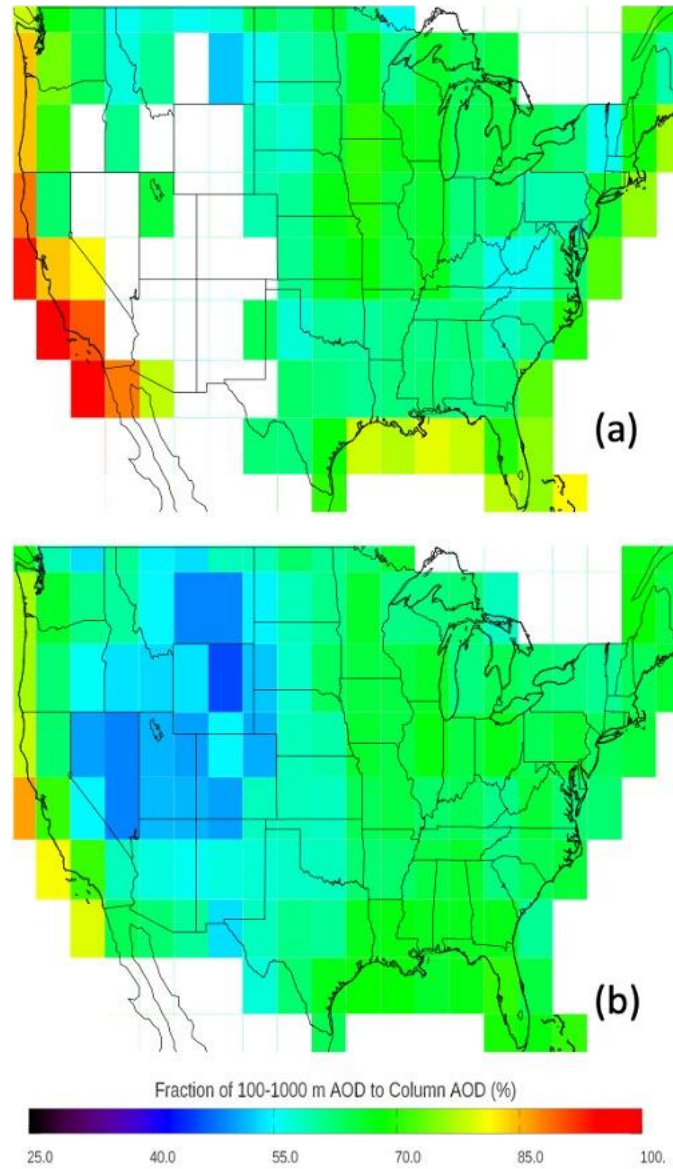


Figure 10. Twelve-year (2007-2018) mean of fraction of 100-1000 m AOD to column AOD (%) on a $3^\circ \times 3^\circ$ grid using (a) daytime and (b) nighttime dust-free CALIOP data.

For the nighttime analysis (Fig. 10b), as during the daytime, percentages around 50-60% are found throughout the CONUS. However, the greater data availability during nighttime over the western CONUS reveals areas with lower percentages (e.g., ~40%-45%). Note that the surface-to-column AOT fractions reported here will be overestimated for those cases for which there is elevated aerosol (e.g., biomass burning smoke) and the CALIOP signal attenuates, due to the near-surface aerosol requirement implemented for this analysis.

Note that we provide yearly mean spatial plots of surface-to-column aerosol representativeness over the CONUS from 2007 to 2018 in Supplemental Fig. 7 (daytime data) and Supplemental Fig. 8 (nighttime data). A more distinct regional spatial pattern is evident in these yearly mean plots, particularly at night, with generally larger percentages in the eastern half of the CONUS compared to the western CONUS. This indicates that aerosol is generally more concentrated near the ground in the eastern CONUS and more vertically diffuse in the western CONUS, a finding consistent with Toth et al. (2014).

The corresponding trends in fraction of 100-1000 m AOT to the total column AOT (% per twelve years) are shown in Supplemental Fig. 9. These were computed with the same procedures and requirements as discussed in Sect. 3.2. As with the previous analyses, the daytime maps show most areas in the western CONUS with no reported trends because of limited data availability in this region (e.g., due to terrain effects and the dust-free requirements), as discussed earlier. While trends of $\pm 10\%$ are found for some grid boxes for both the daytime and nighttime analyses, no clear regional patterns are evident.

4 Conclusions

Using twelve years (2007-2018) of near-surface 532 nm aerosol extinction data from the Cloud-Aerosol Lidar with Orthogonal Polarization (CALIOP) over the contiguous United States (CONUS), particulate matter with aerodynamic diameters smaller than $2.5\ \mu\text{m}$ ($\text{PM}_{2.5}$) concentrations were derived from a bulk-mass-modeling-based algorithm that was recently developed from a proof-of-concept study (Toth et al., 2019). The primary goal of this study is to examine the feasibility of applying this newly developed method to nearly the full CALIOP data

record and study trends of PM_{2.5} from both CALIOP- and ground station-based PM_{2.5} analyses.

We found:

1. The spatial distribution of twelve-year mean PM_{2.5} concentrations derived from near-surface CALIOP aerosol data compares well to that of *in situ* PM_{2.5} measurements collected at U.S. Environmental Protection Agency (EPA) stations. Using twelve years of combined daytime/nighttime near-surface (100-1000 m) CALIOP data, an encouraging relationship is found between CALIOP-derived PM_{2.5} and EPA-observed PM_{2.5} (Deming slope = 0.89; RMSE = 3.42 µg/m³; mean bias = -1.00 µg/m³).
2. The maps of daytime and nighttime CALIOP-derived PM_{2.5} trends over the CONUS suggest near-zero changes in PM_{2.5} concentrations for many areas, but with a noticeable decreasing pattern in PM_{2.5} over the majority of the eastern CONUS (indicating an improvement in air quality over the 2007-2018 time period), a result consistent with EPA-based PM_{2.5}. However, some parts of the western CONUS (e.g., Washington, Oregon, and northern California) show a noticeable increasing pattern in nighttime CALIOP-derived PM_{2.5}, but this is not always consistent with daytime CALIOP and EPA-based PM_{2.5}, likely due to temporal/spatial sampling differences between the datasets.
3. The regionally averaged data over the CONUS (West, Central, Southeast, Northeast) reveals that yearly mean EPA-based PM_{2.5} concentrations more closely match daytime CALIOP-derived PM_{2.5} concentrations than those derived using nighttime CALIOP observations (EPA and daytime PM_{2.5} are larger) for the CONUS and each of the four regions consistently throughout the study period.

Also, agreeable negative trends are found between EPA and CALIOP -based PM_{2.5} concentrations for all regions except the western CONUS, which exhibits positive daytime and nighttime CALIOP PM_{2.5} trends.

4. A seasonal analysis of the EPA and CALIOP PM_{2.5} datasets reveals that over the western CONUS, PM_{2.5} levels trend upward during the more active wildfire season (June through November) but trend downward during other months (December through May). This seasonal pattern demonstrates that the western and eastern CONUS exhibit agreeable negative PM_{2.5} trends over the part of the year for which the western CONUS is not as affected by wildfires, indicative of the air quality regulations implemented in both regions.
5. The assumed mixed layer height in our algorithm is slightly sensitive to the relationship of CALIOP-derived PM_{2.5} to EPA-based PM_{2.5} for the daytime CALIOP analysis but more so for the nighttime CALIOP analysis. Also, the largest daytime and nighttime CALIOP-EPA PM_{2.5} mean biases are found when using only the 0-100 m AGL layer during PM_{2.5} derivation. The poor performance of the algorithm for this layer, and the possible factors impacting it (e.g., surface return contamination), warrant further examination.
6. A noisy spatial distribution of trends in surface-to-column aerosol representativeness, or fraction of 100-1000 m aerosol optical thickness (AOT) to the total column AOT, is exhibited throughout the CONUS.

Accurate monitoring and analysis of surface PM_{2.5} pollution using lidar observations is feasible and has advantages over passive sensor-based methods that use the relationship between AOT and PM_{2.5}. While the passive sensor AOT approach is inherently limited by a column-

integrated perspective of aerosol loading, the distribution of aerosols in the atmospheric column can be obtained from lidar measurements (including the near-surface layers that can be used for PM_{2.5} applications). Also, lidar observations provide valuable PM_{2.5} information at nighttime (not achievable using passive sensor data), insight into the vertical distribution of PM_{2.5} pollution, and a source of validation of air quality models. Still, it is challenging to characterize PM_{2.5} concentrations from space-based lidars due to their lack of swath.

Major efforts that remain in developing more robust lidar-based PM_{2.5} estimates include investigating spatial and temporal representativeness issues, analyzing the impact of varying aerosol types and corresponding mass extinction efficiencies (as well as related assumptions such as the removal of dust), and incorporating the mixed layer/planetary boundary layer height as derived from lidar data or provided by models. This study provides compelling motivation to address these challenges and further examine and improve aerosol extinction-based PM_{2.5} concentration retrievals using current and future space-based lidar observations, including the ATLID instrument (do Carmo et al., 2021) on the Earth Clouds, Aerosol, and Radiation Explorer (EarthCARE; Illingworth et al., 2015) satellite, and the lidar instruments associated with the NASA Atmosphere Observing System (AOS), the satellite mission born out of the Aerosols and Clouds, Convection, and Precipitation (ACCP) pre-formulation study (Stephens et al., 2021).

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

CRediT authorship contribution statement

T.D. Toth: Conceptualization, Methodology, Software, Formal analysis, Investigation, Writing - Original Draft, Writing - Review & Edit, Visualization. **J. Zhang:** Conceptualization, Methodology, Writing - Original Draft, Writing - Review & Edit. **M.A. Vaughan:** Methodology, Writing - Original Draft, Writing - Review & Edit. **J.S. Reid:** Methodology, Writing - Original Draft, Writing - Review & Edit. **J.R. Campbell:** Methodology, Writing - Original Draft, Writing - Review & Edit.

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Figure and Table Captions

Figure 1. Map of the CONUS showing the locations of U.S. EPA stations that report daily PM_{2.5} concentration observations (Parameter Code: 88101) during the study time period (2007-2018). The red lines delineate the boundaries of four regions: West ($\leq -110^\circ$ longitude), Central ($> -110^\circ$ and $\leq -85^\circ$ longitude), Northeast ($> -85^\circ$ longitude and $\geq 40^\circ$ latitude), and Southeast ($> -85^\circ$ longitude and $< 40^\circ$ latitude).

Figure 2. For 2007 to 2018, yearly mean PM_{2.5} concentrations, computed from daily measurements and gridded at $3^\circ \times 3^\circ$ latitude/longitude resolution, from EPA sites across the CONUS.

Figure 3. For 2007 to 2018 over the CONUS, $3^\circ \times 3^\circ$ yearly mean PM_{2.5} concentrations derived from daytime CALIOP near-surface (100-1000 m AGL) aerosol extinction.

Figure 4. For 2007 to 2018 over the CONUS, $3^\circ \times 3^\circ$ yearly mean PM_{2.5} concentrations derived from nighttime CALIOP near-surface (100-1000 m AGL) aerosol extinction.

Figure 5. For 2007-2018 over the CONUS, scatterplots of yearly mean PM_{2.5} concentrations from EPA sites and those derived from collocated near-surface CALIOP aerosol extinction, using (a) daytime, (b) nighttime, and (c) combined daytime and nighttime CALIOP data. Points are color-coded by the number of data points per $1 \mu\text{g}/\text{m}^3$ bin. The dashed and solid lines show the one-to-one lines and Deming regression fits, respectively.

Figure 6. Yearly means of PM_{2.5} concentrations from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP observations for (a) the CONUS and four regions within the CONUS: (b)West, (c) Central, (d) Southeast, and (e) Northeast.

Figure 7. Monthly means of PM_{2.5} concentrations from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP observations for (a) the CONUS and four regions within the CONUS: (b)West, (c) Central, (d) Southeast, and (e) Northeast.

Figure 8. Twelve-year (2007-2018) mean PM_{2.5} concentrations (a) at EPA stations and those derived from (b) daytime, and (c) nighttime, CALIOP near-surface aerosol extinction (gridded at 3° x 3° latitude/longitude resolution). Also shown are the corresponding ratios of (d) daytime to nighttime CALIOP-derived PM_{2.5}, (e) daytime CALIOP PM_{2.5} to EPA PM_{2.5}, and (f) nighttime CALIOP PM_{2.5} to EPA PM_{2.5}, computed for only those grid boxes with available data for each of the analyses in Fig. 8a-c.

Figure 9. Twelve-year (2007-2018) PM_{2.5} concentration trends (a) at EPA stations and those computed from CALIOP measurements for (b) daytime, and (c) nighttime, conditions (gridded at 3° x 3° latitude/longitude resolution). Hatched grid boxes indicate trends that are significant at the 95% confidence interval, calculated using the Mann-Kendall Test.

Figure 10. Twelve-year (2007-2018) mean of fraction of 100-1000 m AOD to column AOD (%) on a 3° x 3° grid using (a) daytime and (b) nighttime dust-free CALIOP data.

1050 Table 1. Twelve-year (2007-2018) trends in PM_{2.5} concentrations ($\mu\text{g}/\text{m}^3$) at EPA stations and
1051 derived from daytime and nighttime CALIOP observations for the CONUS and four regions within
1052 the CONUS: West, Central, Southeast, and Northeast. The trends were computed using the yearly
1053 mean PM_{2.5} concentrations shown in Fig. 5.

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1055 Table 2. The PM_{2.5} trends and their locations for those 3° x 3° grid boxes with statistically
1056 significant (95% CI) daytime or nighttime CALIOP PM_{2.5} trends, and the corresponding EPA
1057 PM_{2.5} trends, as determined from Fig. 6. The corresponding EPA PM_{2.5} trends that are statistically
1058 significant (95% CI) are also indicated.

1059

1060 Table 3. Twelve-year (2007-2018) trends in PM_{2.5} concentrations ($\mu\text{g}/\text{m}^3$) at EPA stations and
1061 derived from daytime and nighttime CALIOP observations for the CONUS and four regions within
1062 the CONUS for each season: December, January, February (DJF), March, April, May (MAM),
1063 June, July, August (JJA), and September, October, November (SON). The trends were computed
1064 using the seasonal mean PM_{2.5} concentrations shown in Supplemental Figs. 3-6.

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1066 Table 4. Results of a sensitivity study varying the height of the assumed mixed layer, including
1067 R², slope computed from Deming regression analysis, and mean bias (CALIOP – EPA; $\mu\text{g}/\text{m}^3$).
1068 This analysis includes the layer nearest the surface (0-100 m).

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1070 Table 5. Results of a sensitivity study varying the height of the assumed mixed layer, including
1071 R², slope computed from Deming regression analysis, and mean bias (CALIOP – EPA; $\mu\text{g}/\text{m}^3$).
1072 This analysis excludes the layer nearest the surface (0-100 m).

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1074 Supplemental Figure 1. For 2007 to 2018 over the CONUS, 3° x 3° yearly mean aerosol extinction
1075 (100-1000 m layer AGL) derived from daytime CALIOP measurements.

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1077 Supplemental Figure 2. For 2007 to 2018 over the CONUS, 3° x 3° yearly mean aerosol extinction
1078 (100-1000 m layer AGL) derived from nighttime CALIOP measurements.

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1080 Supplemental Figure 3. Seasonal means (December through February) of PM_{2.5} concentrations for
1081 each year from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP
1082 observations for (a) the CONUS and four regions within the CONUS: (b) West, (c) Central, (d)
1083 Southeast, and (e) Northeast.

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1085 Supplemental Figure 4. Seasonal means (March through May) of PM_{2.5} concentrations for each
1086 year from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP
1087 observations for (a) the CONUS and four regions within the CONUS: (b) West, (c) Central, (d)
1088 Southeast, and (e) Northeast.

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1090 Supplemental Figure 5. Seasonal means (July through August) of PM_{2.5} concentrations for each
1091 year from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP
1092 observations for (a) the CONUS and four regions within the CONUS: (b) West, (c) Central, (d)
1093 Southeast, and (e) Northeast.

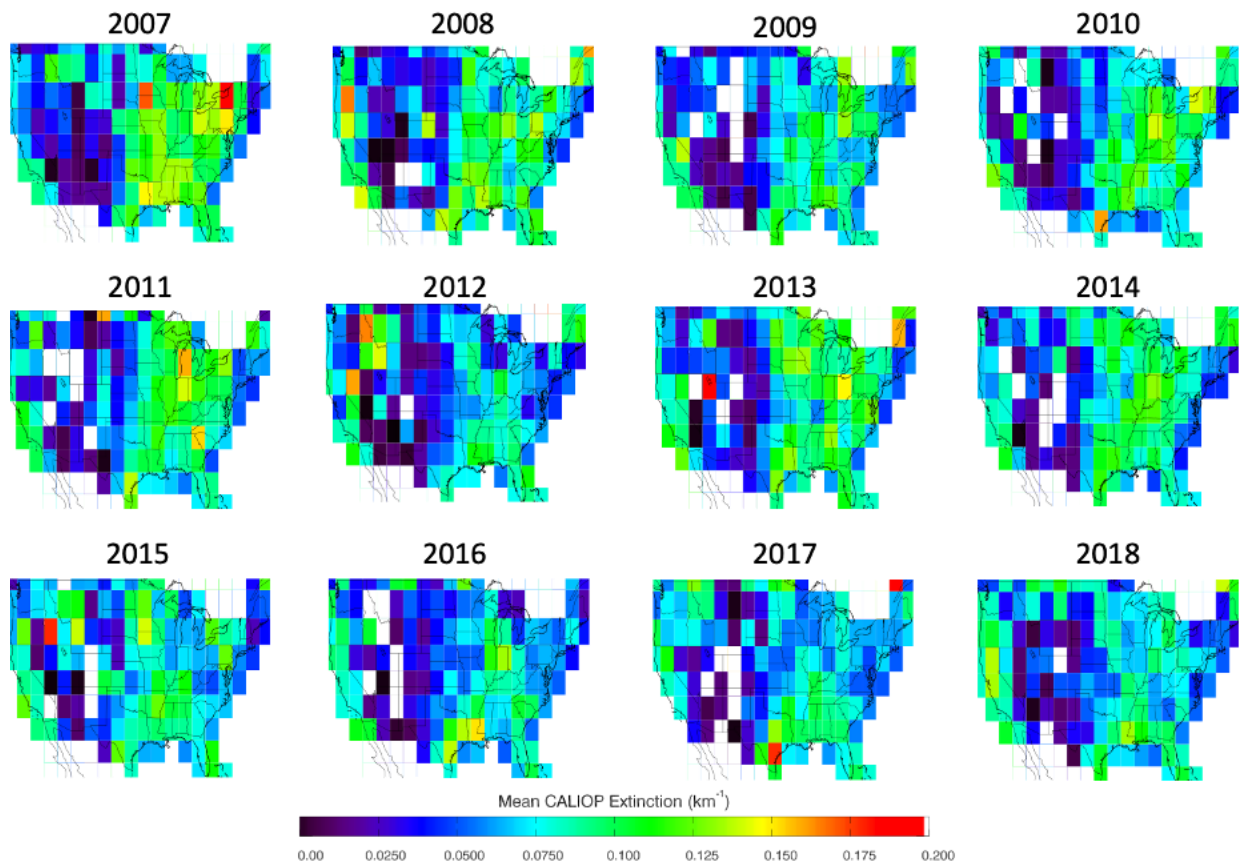
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Supplemental Figure 6. Seasonal means (September through November) of PM_{2.5} concentrations for each year from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP observations for (a) the CONUS and four regions within the CONUS: (b) West, (c) Central, (d) Southeast, and (e) Northeast.

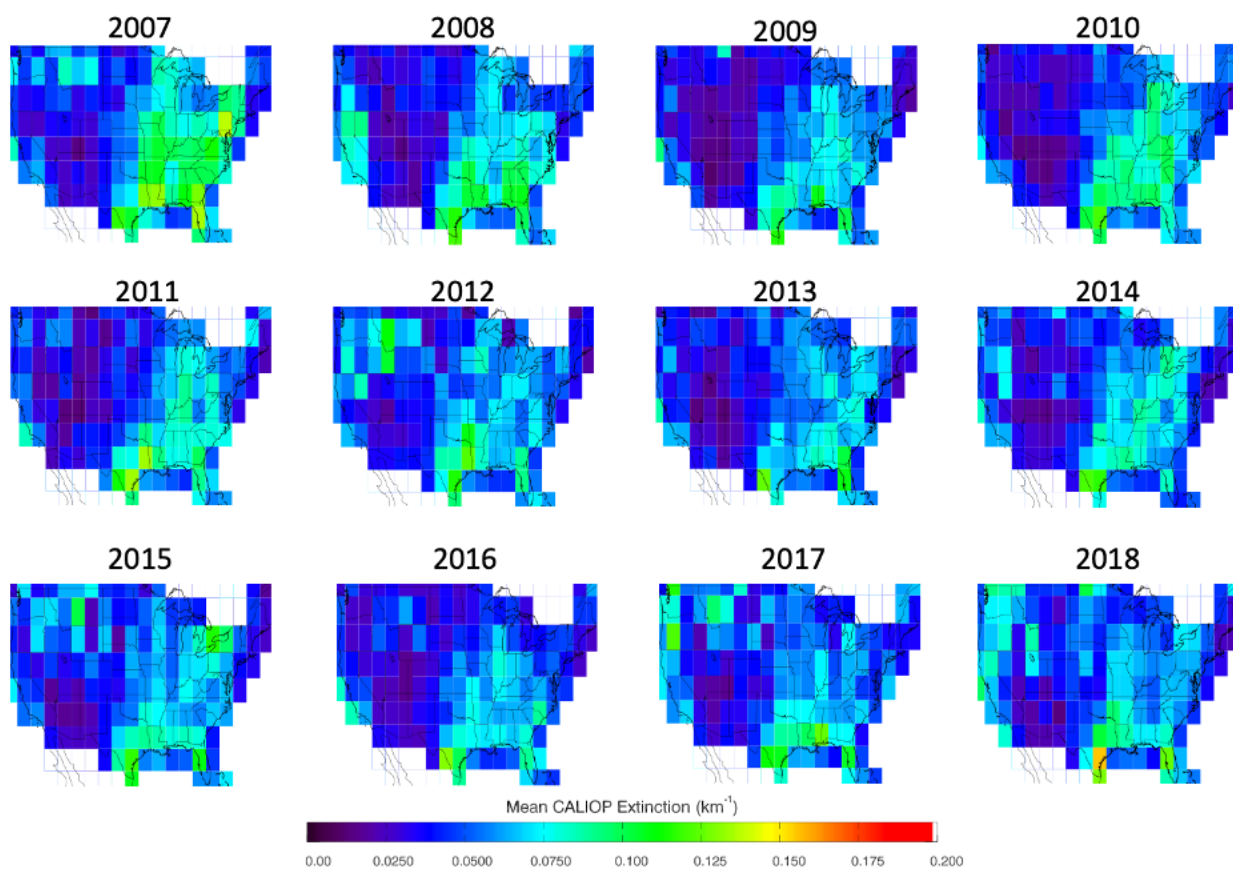
Supplemental Figure 7. For 2007 to 2018 over the CONUS, 3° x 3° yearly mean fraction of 100-1000 m AOD to column AOD (%) as observed by CALIOP during dust-free daytime conditions.

Supplemental Figure 8. For 2007 to 2018 over the CONUS, 3° x 3° yearly mean fraction of 100-1000 m AOD to column AOD (%) as observed by CALIOP during dust-free nighttime conditions.

Supplemental Figure 9. Twelve-year (2007-2018) trends of fraction of 100-1000 m AOD to column AOD (% per twelve years) on a 3° x 3° grid using (a) daytime and (b) nighttime dust-free CALIOP data.

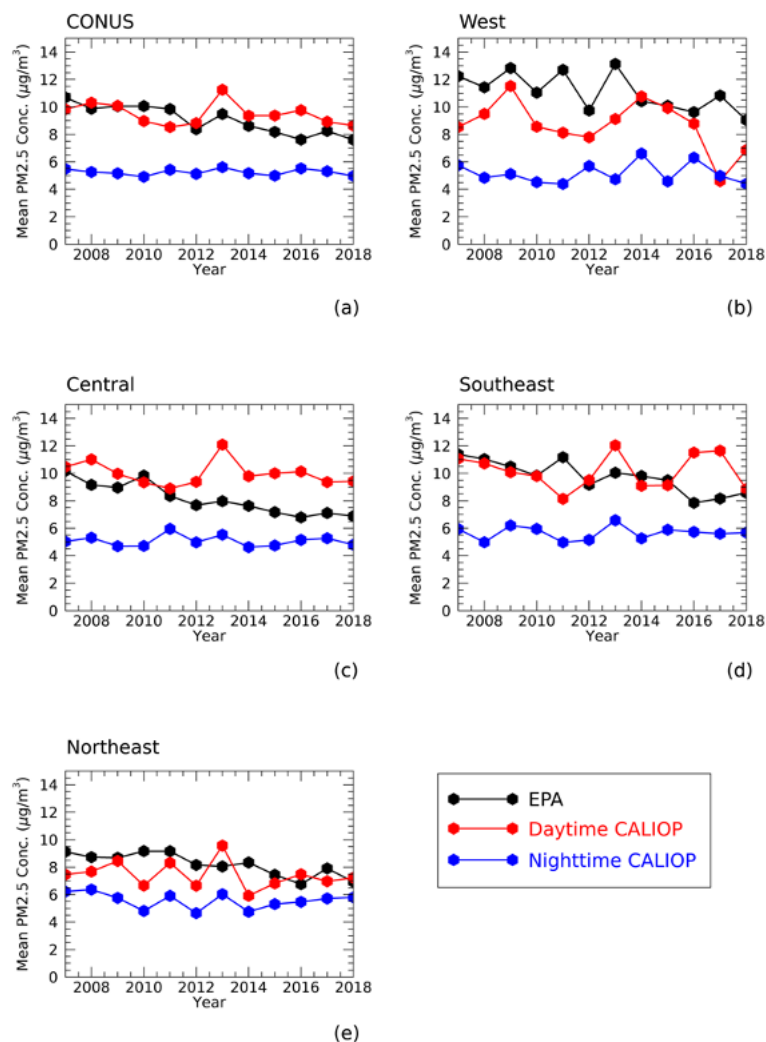


Supplemental Figure 1. For 2007 to 2018 over the CONUS, 3° x 3° yearly mean aerosol extinction (100-1000 m layer AGL) derived from daytime CALIOP measurements.

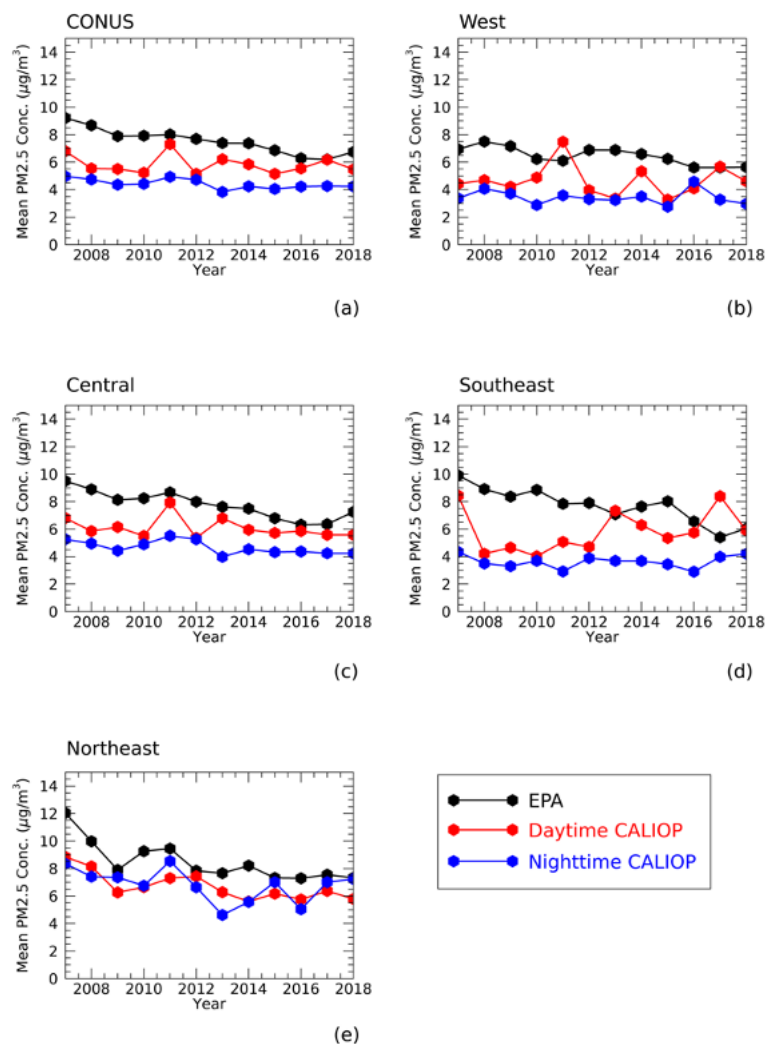


Supplemental Figure 2. For 2007 to 2018 over the CONUS, $3^\circ \times 3^\circ$ yearly mean aerosol extinction (100-1000 m layer AGL) derived from nighttime CALIOP measurements.

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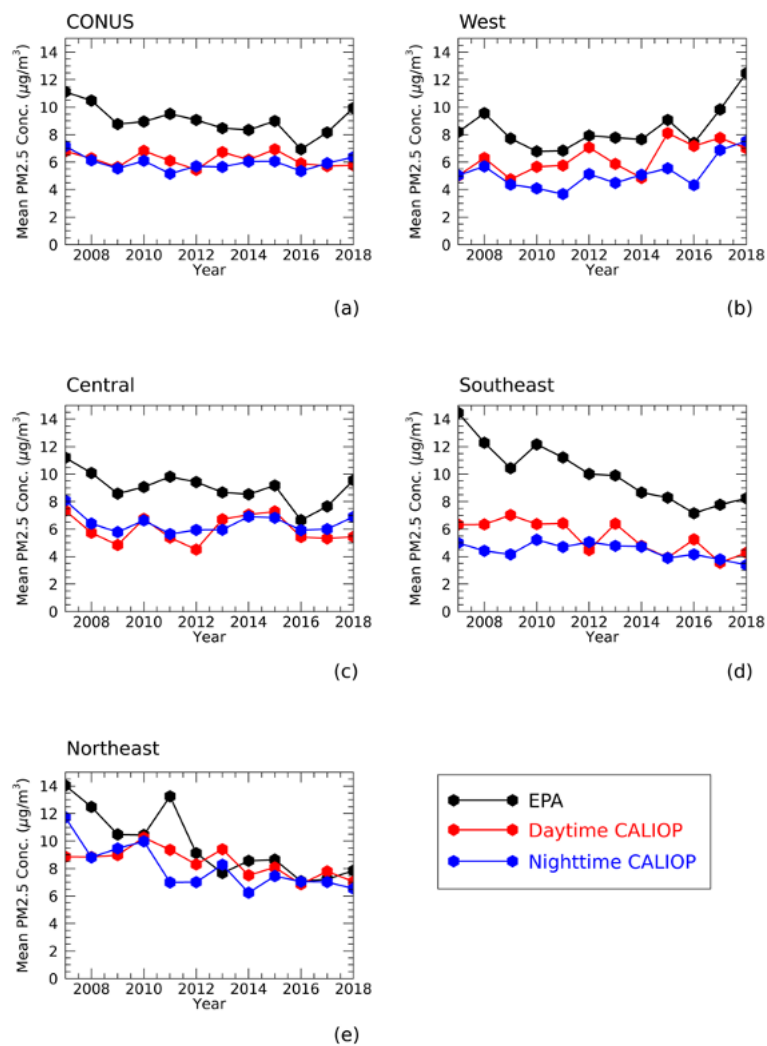


Supplemental Figure 3. Seasonal means (December through February) of PM_{2.5} concentrations for each year from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP observations for (a) the CONUS and four regions within the CONUS: (b) West, (c) Central, (d) Southeast, and (e) Northeast.

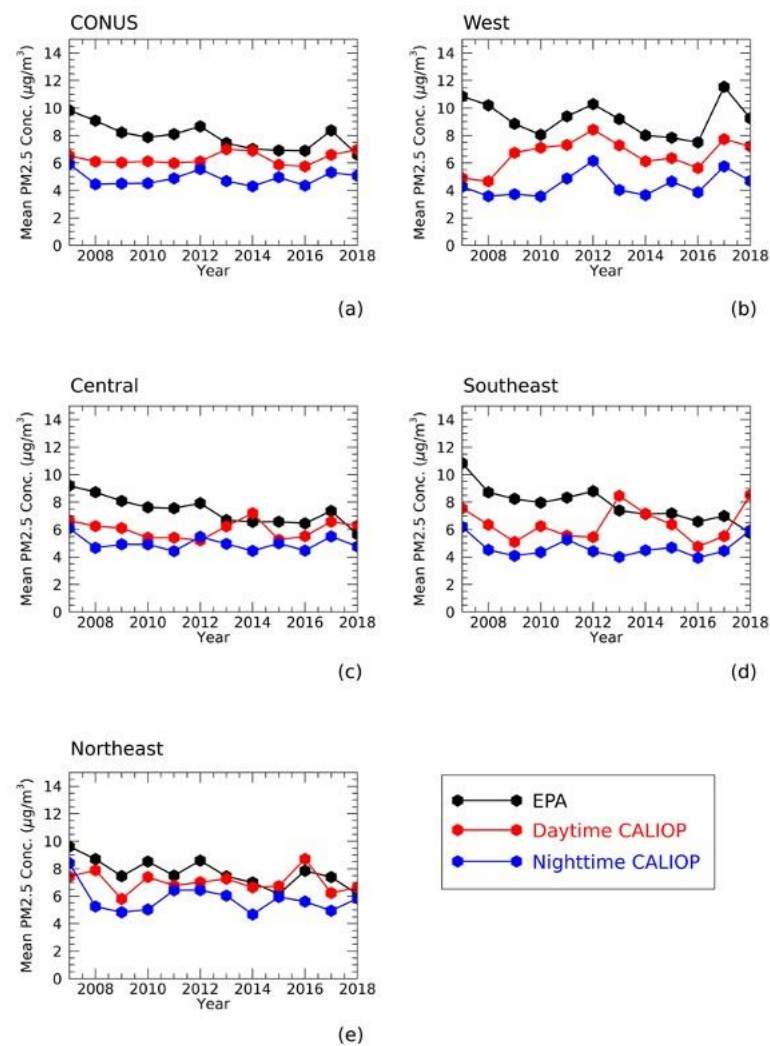


Supplemental Figure 4. Seasonal means (March through May) of PM_{2.5} concentrations for each year from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP observations for (a) the CONUS and four regions within the CONUS: (b) West, (c) Central, (d) Southeast, and (e) Northeast.

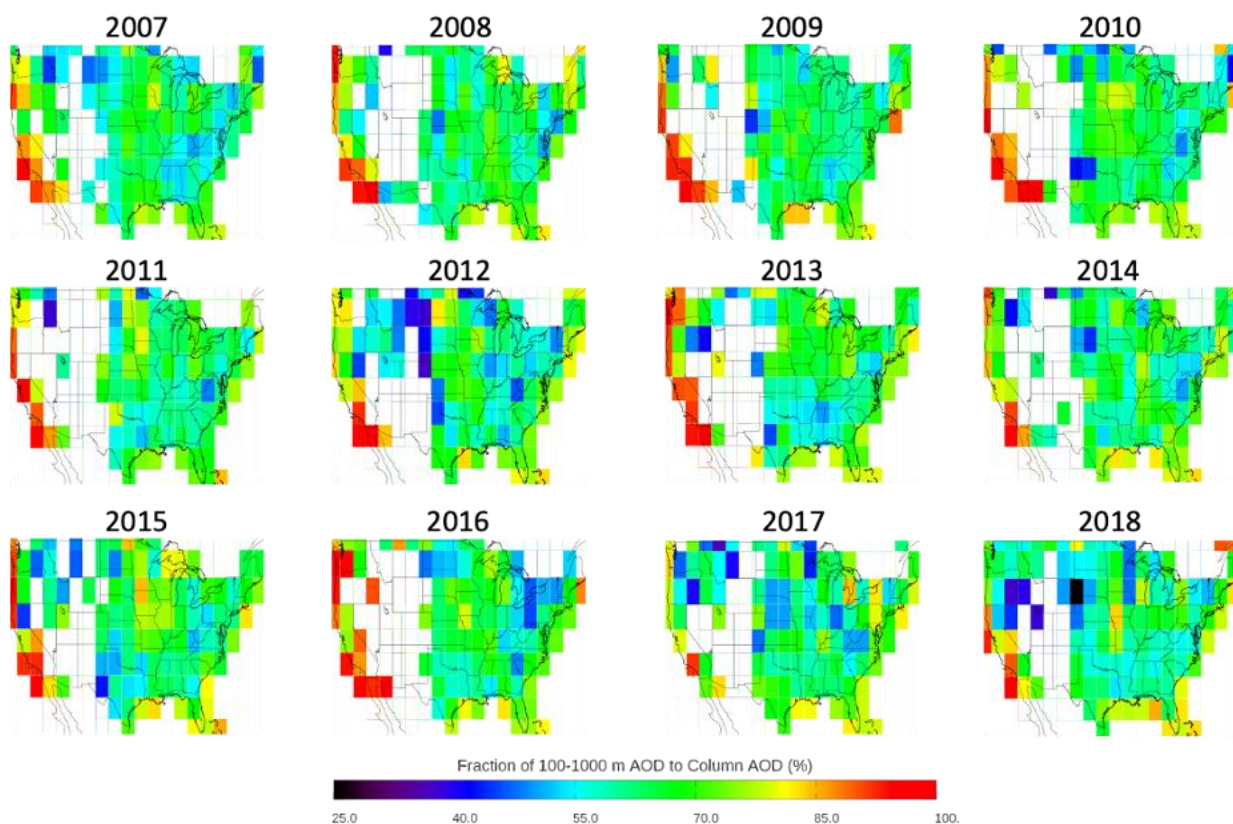
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Supplemental Figure 5. Seasonal means (July through August) of PM_{2.5} concentrations for each year from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP observations for (a) the CONUS and four regions within the CONUS: (b) West, (c) Central, (d) Southeast, and (e) Northeast.

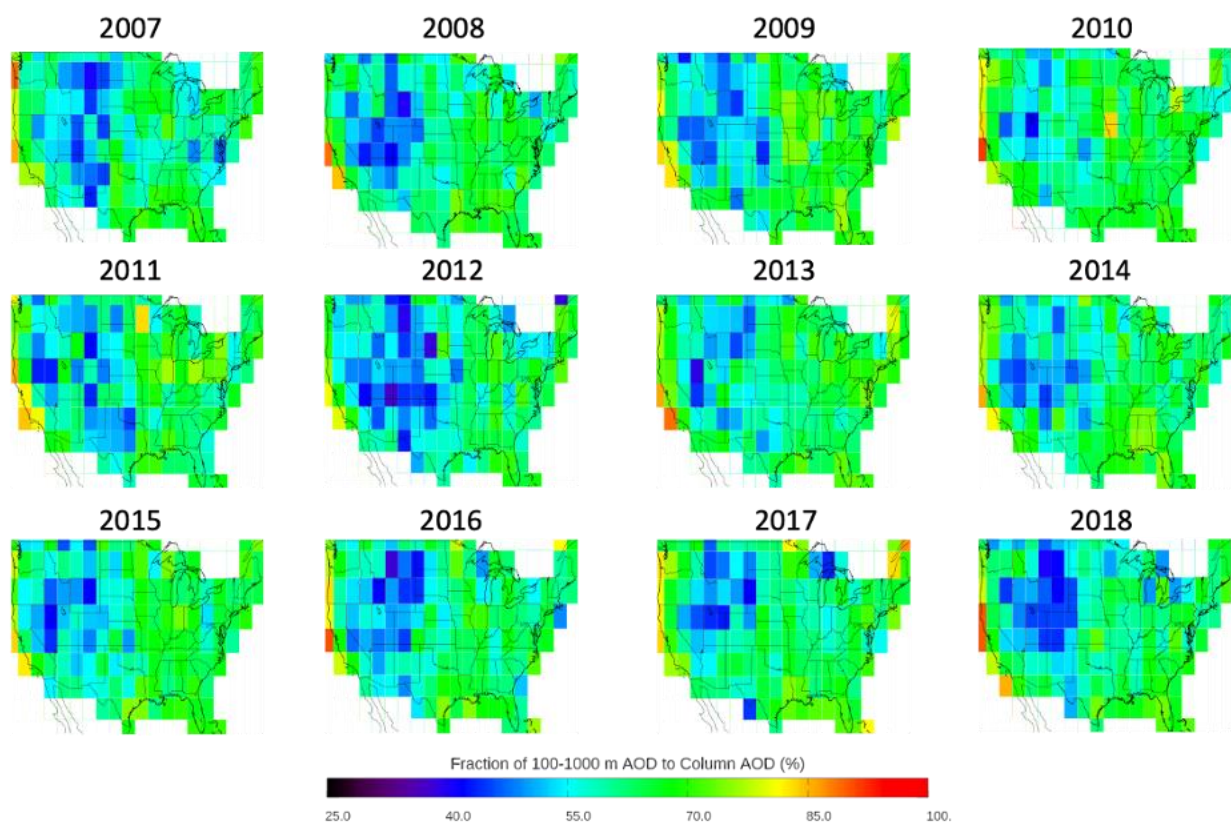


Supplemental Figure 6. Seasonal means (September through November) of PM_{2.5} concentrations for each year from 2007 to 2018 at EPA stations and derived from daytime and nighttime CALIOP observations for (a) the CONUS and four regions within the CONUS: (b) West, (c) Central, (d) Southeast, and (e) Northeast.



Supplemental Figure 7. For 2007 to 2018 over the CONUS, $3^\circ \times 3^\circ$ yearly mean fraction of 100-1000 m AOD to column AOD (%) as observed by CALIOP during dust-free daytime conditions.

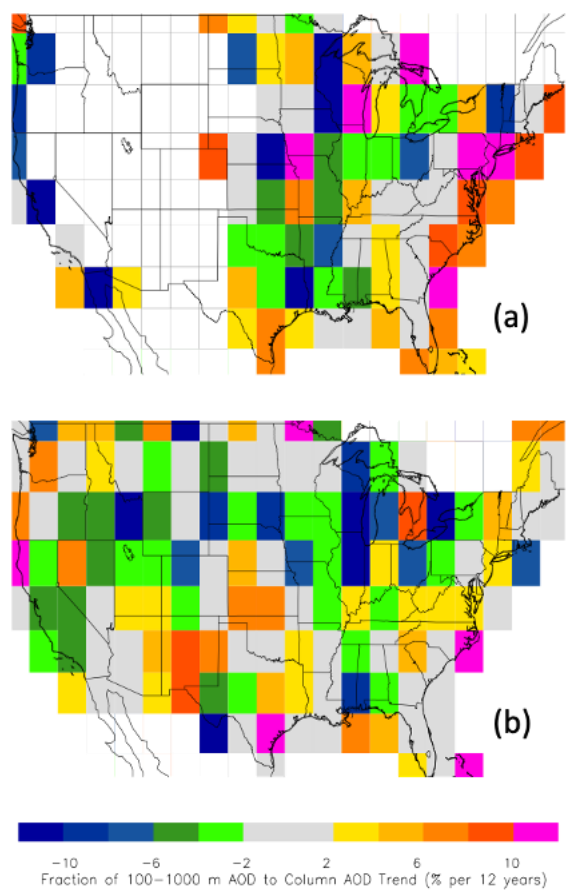
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Supplemental Figure 8. For 2007 to 2018 over the CONUS, 3° x 3° yearly mean fraction of 100-1000 m AOD to column AOD (%) as observed by CALIOP during dust-free nighttime conditions.

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Supplemental Figure 9. Twelve-year (2007-2018) trends of fraction of 100-1000 m AOD to column AOD (% per twelve years) on a 3° x 3° grid using (a) daytime and (b) nighttime dust-free CALIOP data.

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